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ARTICLE

Exposure and behavioral responses of tagged beluga whales (*Delphinapterus leucas*) to ships in the Pacific Arctic

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Abstract

Arctic marine mammals face a multitude of challenges linked to climate change, including increasing anthropogenic noise from ship traffic. The beluga whale (Delphinapterus leucas), a predominately Arctic endemic cetacean, relies heavily on acoustic communication, with documented overlap between their vocalizations and hearing range and ship noise. Some belugas migrate through areas with the highest levels of ship traffic in the Pacific Arctic and exposure to ship noise is highly probable. Here, we document the responses of nine satellite-tagged Eastern Beaufort Sea belugas to encounters with ships in the Beaufort, Chukchi, and Bering Seas during July-December 2018. We report 177 occasions when ships were within 125 km of tagged belugas and quantified changes in lateral and vertical movements to investigate individual behavioral responses to ship approaches within 50 km (n = 23). Belugas' swim speed was negatively correlated with ship distance, showing possible changes in swim speed up to 79 km away. Changes in lateral and vertical movements, indicating disruption of

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KEYWORDS

acoustic disturbance, anthropogenic noise, Arctic, automatic identification system, avoidance behavior, behavioral response, beluga whale, bio-logging tags, cetacean, disturbance threshold, ship traffic

1 | INTRODUCTION

The relatively pristine acoustic marine environment of the Arctic is changing rapidly at least in part as a result of expanding anthropogenic activities (Protection of the Arctic Marine Environment [PAME], 2019). Historically characterized by extensive ice cover, strong winds and low anthropogenic pressure, the Arctic was often a relatively quiet environment as a result of sea ice contributing to extremely low ambient underwater sound levels (median levels [50–1000 Hz band] <75 dB re 1 μ Pa under solid sea ice and ~ 95 dB re 1 μ Pa in the summer open water season sensu Diachok, 1976; Halliday, Barclay, et al., 2021; Insley et al., 2017; Kinda et al., 2013; Roth et al., 2012; Yang & Votaw, 1981). Due to economic, geopolitical, and drastic environmental changes occurring as a result of the pronounced rate of climate change in the region, where warming is two to three times faster than the global mean (Council of Canadian Academics, 2019; Meredith et al., 2019), ship traffic and its subsequent noise footprint have markedly increased over the past few decades (Dawson et al., 2018; Intergovernmental Panel on Climate Change, 2022), and continued increases are projected as trans-Arctic shipping routes become more reliably ice-free (Bennett et al., 2020; Mudryk et al., 2021; Zeng et al., 2020).

The naturally low levels of underwater sound have allowed marine animals to hear anthropogenic noise from farther away than in non-Arctic regions due to reduced acoustic masking by background conditions (Halliday et al., 2020a; Pine et al., 2018). The traditionally lower levels of anthropogenic activity in most areas of the Arctic pair with a lower exposure history to anthropogenic noise for Arctic species compared to non-Arctic species (PAME, 2019). Thus, Arctic marine animals may be more sensitive to anthropogenic noise (e.g., Finley et al., 1990; Halliday et al., 2020a) and noise intensification could have disproportionate effects on Arctic species. As a result, noise impacts on Arctic species may need to be treated as a special case.

For most marine mammals, especially cetaceans, hearing is considered to be their primary sensory modality and they use sound for essential biological functions such as communication, foraging, navigation, and predator avoidance (Tyack, 1986). Underwater ambient noise is generated by both natural (biophony and geophony) and anthropogenic sources. There have been numerous studies documenting negative effects of underwater anthropogenic noise on marine species, with the greatest research focus on marine mammals (e.g., Duarte et al., 2021; PAME, 2019; Southall et al., 2007, 2019). Anthropogenic noise affects marine mammals mainly through acoustic masking or behavioral disturbance. Acoustic masking is defined as the overlap in frequency of anthropogenic signals which interfere with detection of important biological or nonvocal sounds produced by these species (Clark et al., 2009; Payne & Webb, 1971). Disturbance occurs when noise elicits a reaction or a change in behavior (e.g., Finley et al., 1990). However, variability exists across species and among individuals within populations due to a variety of factors, including age, sex, behavioral context, and prior exposure, thereby resulting in variable thresholds of disturbance (see Gomez et al., 2016 for a review; Richardson et al., 1995). Ocean noise from shipping and other maritime activities is now recognized as an acute and chronic, habitat-level stressor (Chou et al., 2021; Duarte et al., 2021; Williams et al., 2020).

Acoustic masking and behavioral disturbance from ship noise (<1 kHz) affect taxa with high hearing sensitivity at lower frequencies (i.e., baleen whales, most pinnipeds, fish, and invertebrates). However, high amplitude ship noise levels do not require high hearing sensitivity at lower frequencies to elicit masking and disturbance. Ship noise can also occur at higher frequencies (1–30 kHz; Veirs et al., 2016), which are especially relevant for odontocete hearing sensitivity (Aguilar Soto et al., 2006; Arveson & Vendittis, 2000; Götz et al., 2009; Vergara et al., 2021). Estimates of detection distances for ship noise vary by frequency, vessel, background noise conditions, and receiving species. The focal species of this study is the beluga whale (*Delphinapterus leucas*), an endemic Arctic odontocete species that has a reduced hearing sensitivity to sounds below 1 kHz (Awbrey et al., 1988; Johnson et al., 1989; Ridgway et al., 2001; White, 1978). However, higher frequency components of ship noise could be audible to belugas at long distances because ship source levels can be relatively high even at above 1 kHz, and belugas hear well at those frequencies (Castellote et al., 2014; Cosens & Dueck, 1993; Mooney et al., 2018, 2020; Popov et al., 2013). Understanding hearing abilities at the species level is essential to determine the effects of noise impacts such as masking, disturbance, and (at high received levels) noise-induced hearing loss.

Belugas generally are gregarious, traveling in pods ranging from a few individuals to hundreds of whales (Jefferson et al., 1993), and are considered to be highly sensitive and vulnerable to ecosystem change (Hauser et al., 2018). In the Pacific Arctic, belugas are an important traditional food source for Inuvialuit and Iňupiat subsistence hunters (Frost & Suydam, 2010; Harwood & Smith, 2002). Belugas in the Pacific Arctic migrate through United States, Canadian, and Russian waters, passing through areas with elevated levels of ship traffic and potentially are exposed annually to a high number of acoustic disturbance events (Halliday, Pine, et al., 2021). A recent review of underwater noise and Arctic marine mammals (Halliday et al., 2020a) showed that belugas can be disturbed by ship-related noise and temporarily displaced. A few studies have examined the impacts of underwater noise on the behavior of wild belugas, and in some cases noise from outboard motors, icebreakers, tugs, barges, seismic air guns, and drilling evoked an avoidance/startle (i.e., flee) response at varying received levels (Blevins, 2015; Cosens & Dueck, 1988; Finley et al., 1990; Fraker, 1977, 1978; Koski et al., 1995; Krasnova et al., 2009; Miller et al., 2005; Richardson et al., 1995; Stewart et al., 1982, 1983). Belugas showed the strongest reported reactions to icebreaker ships with avoidance responses occurring at 35-50 km distance from the icebreaker where received noise levels ranged from 94 to 105 dB re 1 µPa [20-1,000 Hz band] (Cosens & Dueck, 1993; Finley et al., 1990). The events involving icebreaker ships resulted in the displacement of all belugas from the region for periods of 1-2 days and by up to 80 km (Finley et al., 1990; LGL & Greeneridge, 1986). Further, Miller et al. (2005) suggested that belugas avoided areas with ships conducting seismic operations by distances of 10-20 km. A recent study by Halliday et al. (2019) showed that beluga vocalizations decreased when ships traveled within 5 km of a moored acoustic recorder. This reduction was either caused by belugas decreasing their calling rates or fleeing the area in response to the ship; the latter is supported from observations by Inuvialuit (Halliday et al., 2019). In addition, noise from manned aircraft and unmanned aerial vehicles (UAVs) flown near the water in some cases has caused a flee response in belugas in the Arctic (e.g., Fraker, 1978; Palomino González et al., 2021; Patenaude et al., 2002), further supporting beluga heightened sensitivity to anthropogenic noise compared to other species of marine mammals.

Avoidance behaviors by animals exposed to threatening stimuli are generally grouped into two categories: "fight or flight" reactions and "freeze" reactions (Gabrielsen & Smith, 1995; Roelofs, 2017). Fleeing mammals display physiological responses characteristic of exercise (i.e., tachycardia, increased metabolic rate, increased swim speed; Ford & Reeves, 2008), while mammals demonstrating freeze responses experience bradycardia and metabolic slow-downs (Roelofs, 2017; Steen et al., 1988). However, a paradoxical physiological response was reported in tagged narwhal displaying simultaneous bradycardia with increased fluke stroke and respiration rates (Williams et al., 2022). Previous research suggests that belugas are more prone to exhibit a flee response to a perceived threat compared to a fight (e.g., aggressive approach) or freeze response. The flee response of belugas has been described as large herds undertaking long dives, where pod integrity breaks down into small, scattered groups and diving becomes

asynchronous (Blevins, 2015; Finley et al., 1990; Palomino González et al., 2021). Krasnova et al. (2009) found that female belugas with calves are usually the first group members to flee from anticipated danger. Fleeing also causes the individual or group to cease its current behavioral state (e.g., foraging, nursing, resting, transiting), thereby disrupting important daily activities. Interruptions of behavior, even if short-term, have the potential to negatively impact individual belugas if they are repeated (Tyack, 2009). Thus, while fleeing can afford survival from a presumed threat, it can have short- and longer-term impacts on an individual's fitness, which could lead to a reduction in the overall health of individuals and a population. Previous studies that examined beluga reactions to noise describe diving as a common avoidance response but lacked the technology to quantitatively assess changes in dive behavior paired with surface observations. Modern telemetry methods using animal-borne tags allow researchers to capture the full 3-dimensional movements of marine mammals to provide a more thorough examination of behavioral responses to disturbance events (Hussey et al., 2015).

The current study reports location data from satellite-linked time and depth tags attached to male Eastern Beaufort Sea (EBS) belugas in conjunction with ship location data, and summarizes the number of ships encountered by tagged individuals. We then assess data on beluga lateral surface movements, and where possible, dive behavior including identified dive types associated with certain functional behaviors (e.g., foraging, travel, and recovery; Storrie et al., 2022) to characterize potential behavioral responses to vessels. The scope of this study encompasses beluga behavioral responses from near-continuous monitoring over a vast area of the Pacific Arctic including both coastal and offshore waters in three marginal seas. Given the logistical difficulties in performing a controlled vessel-disturbance study involving belugas in the Arctic, and the cultural and ecological importance of this susceptible species, the encounters presented here provide evidence about how belugas may respond to ship noise, as well as information useful in defining hypotheses for future testing to address this issue.

2 | MATERIALS AND METHODS

2.1 | Study area and focal animals

This study focuses on the Pacific Arctic region, from the northern Bering Sea to the Amundsen Gulf in the western Canadian Arctic, and includes the Bering, Chukchi, and Beaufort Seas (Figure 1). During July 3–12, 2018, 10 adult male belugas from the EBS stock were instrumented with back-mounted tags using a live capture method (Orr et al., 2001) at Hendrickson Island within Kugmallit Bay of the Mackenzie River estuary, Northwest Territories, Canada (Table 1; Storrie et al., 2022). One beluga's (LC2018#9) tag transmitted for only approximately 1 week and consequently those data were excluded from the following analyses. For the remaining whales, three were fitted with SMRU CTD-SRDL tags (Sea Mammal Research Unit, University of St. Andrews, Scotland) and six were fitted with SPLASH10-F-238 tags (Wildlife Computers Inc., Redmond, WA; Table 1). Both tag types recorded Argos and Fastloc Global Positioning System (GPS) satellite-derived locations and contained time depth recorders (TDRs) for assessing time series dive behavior. Data for the current study were constrained to satellite locations and diving behavior obtained from these nine tagged belugas during the period July-December 2018 (Figure 1a).

Argos satellite tags have been widely used to track large-scale animal movements; however, tags provide relatively coarse quality locations typically with an accuracy of several hundred meters to several kilometers. Each Argos location is assigned a location accuracy designated by a number or letter dependent on the number of orbiting satellites via which the transmission is received, among other factors (https://www.argos-system.org; accuracy categories: GPS [<100 m], 3 [<250 m], 2 [250–500 m], 1 [500–1,500 m], 0 [>1,500 m], A and B [unbounded], Z [invalid]). Consideration of these location accuracies is important when examining fine scale encounters such as individual animals with ships or conspecifics. A Fastloc GPS receiver on a tag can take a snapshot of up to 10 GPS satellites to provide location accuracies of 10s of meters that are then transmitted via standard Argos transmissions (Dujon

located within the Exclusive Economic Zones (EEZs) of Canada, the United States and a portion of Russia shown in (b) as the region inside of the dashed gray line. System: NAD 1983 2011 Alaska Central Meridian: 154'00'W NAD 1983 2011 Aeridian: 154°00° System: 1 Beluga whale satellite tag locations (July-December 2018)

Ξ

380 570 760

190

AIS ship tracks (July-December 2018)

570 760

100 380

LC2018 #6 LC2018 #7

LC2018#1 LC2018 #2 LC2018 #3 1.C2018#4 LC2018#5 LC2018#8 LC2018

Vhale ID

lonth

Arctic Ocean

(q)

(a)



TABLE 1 Summary	y of tagged belugas and their numbers of known encounters with ships within a radius of 125 km during July-December 2018. Tag type "Splash"
represents a SPLASH1	10-F-238 tag and "SMRU" represents a SMRU CTD-SRDL tag. Ship encounters denotes the closest distance of approach to a given ship on a given day
after removal of duplic	cates when an encounter spanned past midnight. The three closest points of approach (CPA) to ships as well as the average (± SD) CPA distance are
provided per whale. A	w maximum time difference of 1 hr between the paired whale location and AIS ship location was used to calculate the CPAs. Beluga LC2018#5 was not
known to encounter a	any ships within a radius of 125 km in this study.

	h (km)	Avg (SD)	79.6 (± 30.7)	75.2 (± 39.4)	86.9 (± 22.4)	81.2 (± 30.9)	Ι	86.7 (± 27.4)	87.4 (± 38.3)	84.9 (± 24.8)	97.1 (± 19.3)
	f approac	3rd	57.3	13.1	7.7.	40.8	I	20.9	I	66.9	90.4
	t point of	2nd	54.8	12.6	60.7	35.1	I	14.3	114.5	61.1	89.7
	Closes	1st	43.8	6.8	55.0	24.4	I	13.4	60.3	25.2	72.3
er 2018		Total	6	57	ω	15	0	71	2	13	5
-Decembe		101- 125	ო	22	ო	4	0	28	1	4	7
tered July-	(km)	76- 100	0	10	ო	9	0	21	0	9	2
s encount	ance bins	51- 75	5	~	7	7	0	14	7	7	7
Ship	Dista	ο _δ	1	18	0	ო	0	ω	0	1	0
		Tag type	Splash	Splash	Splash	Splash	SMRU	Splash	SMRU	Splash	SMRU
	Duration	(days)	182.7	349.9	161.2	334	17.1	355.7	96.1	162.2	46.7
		Tag off	Jan 2, 2019	Jun 19, 2019	Dec 15, 2018	Jun 7, 2019	Jul 25, 2018	Jun 29, 2019	Oct 13, 2018	Dec 19, 2018	Aug 28, 2018
		Tag on	Jul 3, 2018	Jul 4, 2018	Jul 6, 2018	Jul 8, 2018	Jul 8, 2018	Jul 8, 2018	Jul 9, 2018	Jul 9, 2018	Jul 12, 2018
	Length	(cm)	420	470	406	444	419	440	370	425	434
		Tag ID	174965	174967	174962	174963	175284	174966	175278	174969	175282
		Whale ID	LC2018#1	LC2018#2	LC2018#3	LC2018#4	LC2018#5	LC2018#6	LC2018#7	LC2018#8	LC2018#10

et al., 2014). Although Argos derived locations are coarser than those of Fastloc GPS, they require less energy (i.e., battery power) and typically can be determined for a longer duration.

In the current study, locational information was used from both Argos and Fastloc GPS derived locations, preferentially using Fastloc GPS locations when available. Tags were programmed with an Argos transmission rate of 25 s, and to collect Fastloc GPS locations every 7–30 min; however, temporal gaps between satellite-derived locations were often longer than this as a result of transmissions only occurring when the animal surfaces and tag programming (e.g., fewer transmissions from October or November onwards, see Storrie et al., 2022).

We improved location estimates for each animal using a continuous-time Correlated Random Walk (CRW) model developed by Johnson et al. (2008) and implemented in package "crawl," version 2.2/1 (Johnson & London, 2018), in R (R Core Team, 2021). CRW models are limited by the assumption that the errors follow a normal distribution. In general, Argos location data are close to normal except for the presence of extreme outliers. The CRW algorithm performs poorly when Argos position error greatly exceeds measured position values and CRW model priors. Consequently, extreme outliers were removed by passing the raw location data through the sdafilter in the R package "argosfilter" (Freitas, 2012; Freitas et al., 2008). We used the default speed threshold of 2 m/s, which is greater than the maximum published speed for belugas of 1.78 m/s (Richard et al., 2001). The filter can also be specified to remove outliers which create acute angles in the path of movement (i.e., "spikes"). We used default values for specifying the angular components of the filter; specifically, we removed values that formed angles <15° when they were >2.5 km from the previous location and angles $<25^{\circ}$ when they were >5 km from the prior location. Furthermore, the estimated locations and tracklines were visually verified to ensure the data were not being "pulled" towards low-quality locations that marginally passed the filter. Bayesian state-space switching models do not require prefiltering of Argos locations; however, these are discrete time-step models which typically only estimate a location 1-2 times per day. Those intervals are not frequent enough to pair with ship data to assess whale behavioral responses. Hence, it was necessary to use a continuous-time movement model in this study.

The CRW model treats movement as a velocity process with two parameters, β , the autocorrelation in velocity and σ , the variation in velocity. Location error was assumed to be normally distributed with a mean of 0 and a standard deviation equal to that declared by the system operator, Collecte Localisation Satellites (CLS), for least-squares location classes GPS, 3, 2, and 1 (CLS, 2016). We treated error for the remaining three location classes as parameters to be estimated and fitted them to half normal distributions with semi-informative priors. Locations with classes 0, A, and B should have more error than those with a class of 1 (SD = 1,500 m). Hence, our half normal distributions had a lower bound of 1,500 m. Using data from Vincent et al. (2002), our priors had a mean error of 1,500 m and a standard deviation of 5,000 m for location classes 0 and A, and 7,500 m for location class B. We also set a Laplace prior (double exponential) for β and σ . The Laplace prior had a mean of 3 and a variance of 0.5 on a natural log scale, which is approximately the value of β and σ observed for most species. Note that this is only significant when tracks have few location data. This is the same model and error estimation used for belugas in Citta et al. (2018, 2020). We used the model to better estimate beluga location at preexisting Argos locations; we did not use the model to predict where a beluga might be located between preexisting Argos locations. In effect, the model was only used to reduce location error. The CRW modeled whale locations, hereafter referred to as "whale locations."

2.2 | Spatial and temporal analysis of belugas and ships

The main shipping season for the study region occurs from July to October during the predominantly ice-free period, but can extend to November and December near the Bering Strait. Ship tracks from July to December 2018 were derived from preprocessed satellite Automatic Identification System (AIS) data (exactEarth Ltd., Cambridge, ON, Canada) in the Pacific Arctic spanning from the Chukchi Sea and Bering Strait in the west to the Amundsen Gulf in the east (Figure 1b). AIS transponders transmit signals which show the geographic coordinates and other information about individual ships at regular intervals (e.g., every 2–120 s, depending on ship behavior and context), and these

signals can be received by dedicated satellites and land-based receivers. Internationally, only a subset of vessels are required to carry AIS transponders, specifically all ships \geq 300 gross tonnes on international voyages, cargo ships \geq 500 gross tonnes on domestic voyages, and all passenger ships with >12 passengers (International Maritime Organization, 2014). Other ship types such as barges, tugs, recreational vessels, and research ships are not required to carry AIS voluntary transponders; however, many of these ships use AIS transponders for safety reasons, particularly in the Arctic. Vessel traffic in the Pacific Arctic consists of a variety of vessels including bulk carriers, community supply vessels (barges and tugs), container ships, cruise ships, government icebreakers and research vessels, tankers, military vessels, seismic survey vessels, recreational vessels, and local community traffic for subsistence activities (see Dawson et al., 2018).

We used satellite AIS data to calculate the number of unique vessels that encountered individual tagged belugas in the study area. We acknowledge this may be a conservative estimate if additional ships were present and not transmitting AIS data. When a ship track overlapped in time and space with a tagged beluga, AIS data were used to calculate the range of distances between the ship and the whale including the closest point of approach (CPA) during encounters.

Several steps were taken to calculate the number of unique vessels that encountered individual tagged whales as well as the CPA for each encounter. A spatial and temporal analysis was completed in ArcGIS using ArcMap 10.8 (Environmental Systems Research Institute, Redlands, CA). The first step of the analysis created a buffer radius of 125 km around each individual satellite-derived whale location acquired during July-December 2018. We selected a 125 km radius as modeling has suggested that ship noise can be greater than ambient levels at distances over 100 km in the region (Halliday et al., 2017). Ship noise can theoretically contribute to ambient sound >125 km away, although it is difficult to quantify at this distance (Aulanier et al., 2017). Next, we separated whale location and AIS ship location data by "day" and paired these data sets in space and time. Individual whale and ship data from the same day occurring within a radius of 125 km were then extracted and designated as encounter events. This process was completed separately for each individual whale to ensure that all possible encounters between whales and ships were included in the analysis. Next, the derived paired whale and ship locations within each event were sorted by time to generate time series of consecutive potential encounters while retaining all underlying data and geographic positional information. The maximum allowed delta time between a paired whale and AIS location was 1 hr and the majority had a delta time of <3 min. The "Points to Line" tool in ArcMap 10.8 was used to calculate the distance (meters) between the closest aligned whale and AIS locations as they approached each other within each event. For each encounter occurring within a radius of 125 km on a given day, the CPA was calculated as the shortest distance observed between the whale and ship.

In order to quantify if multiple ships came within 125 km of an individual whale during an encounter event, the CPAs from each whale-ship encounter were compiled and sorted by date and time. This also allowed manual removal of duplicate encounters that resulted from an encounter spanning past midnight where it was included twice based on the original subsetting of data by "day." Once duplicates were removed, all encounters were sorted by the CPA distance and summarized by whale ID (Table 1).

At a finer scale, we chose to investigate all encounter events between whales and ships with a CPA \leq 50 km (Table 2) based on the findings of Finley et al. (1990) where belugas showed strong avoidance reactions to ships approaching at distances of 35–50 km. For each encounter event with a CPA \leq 50 km, approximately 72 hr of consecutive whale locations and AIS ship locations centered around the CPA were extracted. To examine whale behavior in the theoretical absence of vessels, a control period (24 hr) containing no known ships within 125 km was identified from within the same week for each encounter event. Where possible, this control period was taken from the 24 hr before or after a ship came within 125 km of the whale. Whale locations were extracted for each control period and control dates were used only once (Table 2). For individual whales that came within 125 km of multiple ships on the same day, all unique cases of ship and whale locations were aligned in time to assess each ship's location from the whale's perspective to identify if and when more than one ship was located within 50 km of the whale. For

TABLE 2 Summary of 23 encounters with the closest point of approach (CPA) < 50 km distance between a ship and a tagged beluga. Under the column "Behavior," "NLR"
stands for no lateral response, "A" stands for potential avoidance behavior, "UND" stands for undetermined behavior, and "DD" stands for data deficient. Integer values in the
"Behavior" column correspond to the encounter description in the Results section. The estimated broadband received level (RL) in dB re 1µPa of ship noise (maximum in water
column) is provided for the CPA, and delta time (seconds) is the time difference between the paired whale location and AIS ship location used to calculate the CPA. Ship speed
(knots) is provided for the CPA. For encounters with multiple ships, the delta time, CPA distance and ship speed are provided for the closest ship which is listed first under
"Ships present," and RL is a combined estimate for all ships present.

Whale ID	Date	Time (UTC)	Latitude	Longitude	Ships present	CPA (km)	∆ time (s)	RL (dB)	Ship speed (knots)	Behavior	Control date
LC2018#1	Dec 28, 2018	02:03	63°45′01.22″N	173° 58' 00.80''W	Boris Sokolov	43.8	103	103.4	8.2	DD	Dec 26, 2018
LC2018#2	Aug 7, 2018	22:27	70°15′22.06″N	126° 34′ 54.79″W	Frosti + Sir Wilfrid Laurier	19.1	43	97.8	5.5	A 1	Aug 5, 2018
LC2018#2	Aug 9, 2018	09:31	69°58′23.66″N	131°21′17.15″W	Fathom Wave + Kelly Ovayuak	13.9	16	158.7	4.4	UND 1	Aug 4, 2018
LC2018#2	Aug 18, 2018	02:34	69°58'13.74"N	120° 13' 57.19''W	Sir Wilfrid Laurier	23.6	533	113.4	11.1	NLR 1	Aug 16, 2018
LC2018#2	Aug 10, 2018	14:58	69°32'04.43"N	118° 14' 17.54''W	Kelly Ovayuak	12.6	61	126.5	10.0	A 2	Aug 14, 2018
LC2018#2	Aug 22, 2018	10:24	69°59′36.60″N	118° 28' 41.27″W	Sir Wilfrid Laurier + BBC Oregon	39.4	34	119.0	11.1	UND 2	Aug 30, 2018
LC2018#2	Aug 25, 2018	03:48	70°18′49.80″N	122° 42′ 25.66″W	Frosti + Sir Wilfrid Laurier + High Progress	13.1	6	118.9	9.3	A 3	Aug 29, 2018
LC2018#2	Sep 22, 2018	20:23	70°41′38.53″N	141° 54′ 02.49′′W	Frosti + David Thompson	25.5	T	108.3	9.6	UND 3	Sep 21, 2018
LC2018#2	Nov 5, 2018	04:49	69°46′58.14″N	179°36'18.64"E	Andrey Pervozvanniy	6.8	65	123.5	12.4	UND 4	Nov 3, 2018
LC2018#2	Nov 8, 2018	04:56	68°02′51.89″N	175° 40' 24.25''W	Arkadiy Chernyshev	46.7	132	104.0	11.4	NLR 2	Nov 6, 2018
LC2018#2	Nov 12, 2018	04:21	66°53'25.67"/N	171°39′21.46′′W	St. Confidence	44.5	2281	97.3	8.9	DD	Nov 11, 2018
LC2018#2	Nov 19, 2018	03:27	66°41′12.12″N	171°21′38.89″W	Georgiy Brusilov	45.8	12	126.1	18.1	DD	Nov 18, 2018
LC2018#2	Nov 22, 2018	20:37	66°26′13.62″N	169° 58′ 52.92″W	Rudolf Samoylovich	19.8	26	123.5	13.3	UND 5	Nov 29, 2018
LC2018#4	Oct 11, 2018	23:18	71°37′11.01″N	174°58′54.55″W	Vladimir Vize	24.4	9	126.2	15.6	9 OND 6	Oct 10, 2018
LC2018#4	Nov 1, 2018	10:49	68°24′55.04″N	174°02'03.65"W	Nordic Olympic	35.1	1	112.8	11.5	UND 7	Oct 29, 2018
LC2018#4	Nov 8, 2018	07:27	68°13'00.54"N	173°05'02.49''W	Arkadiy Chernyshev	40.8	48	104.8	11.2	A 4	Nov 7, 2018

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TABLE 2	(Continued)										
Whale ID	Date	Time (UTC)	Latitude	Longitude	Ships present	CPA (km)	Δ time (s)	RL (dB)	Ship speed (knots)	Behavior	Control date
LC2018#6	Aug 12, 2018	06:32	70°28′10.80″N	123° 16′ 57.56″W	Fathom Wave + Sir Wilfrid Laurier	43.1	15	98.9	6.0	A 5	Aug 13, 2018
LC2018#6	Aug 24, 2018	04:33	71°14'04.27"N	133° 51' 51.36"W	High Progress	39.3	5	112.1	12.1	NLR 3	Aug 23, 2018
LC2018#6	Aug 29, 2018	09:15	70°34'00.77"N	140° 40' 58.20''W	David Thompson	20.9	16	100.2	9.4	A 6	Aug 28, 2018
LC2018#6	Sep 12, 2018	08:44	72°25'06.22"/N	156° 28' 08.20''W	Sikuliaq + High Progress	14.3	6	111.7	9.5	A 7	Sep 14, 2018
LC2018#6	Nov 4, 2018	00:30	67°50'33.88"N	173° 45' 05.99''W	Rubin	41.3	35	116.8	8.7	DD	Nov 3, 2018
LC2018#6	Nov 19, 2018	06:16	67°22′27.50″N	171°27′59.54″W	Georgiy Brusilov	13.4	11	133.1	17.7	A 8	Nov 17, 2018
LC2018#8	Nov 23, 2018	03:07	69°06′22.72″N	169° 18′ 43.03″W	Mirai	25.2	76	118.0	11.6	DD	Nov 21, 2018

each encounter event with a CPA \leq 50 km, an animation was created in ArcMap 10.8 by pairing the whale and ship AIS locations by time using the animation toolbar (see Supplementary Materials). If multiple ships were present in an encounter event, they were included in a single animation. Animations provided a moving visual perspective of the encounter from a bird's eye view and were used in part to identify evidence of lateral responsive movement by a whale in response to ships during events.

2.3 | Estimation of received level of ship noise

In this study, ship noise refers to sound created by propeller cavitation, engine noise, and other sounds accompanying normal ship operations. We do not address noise from air guns, sonar, or other noises associated with military or resource exploration activities. To estimate the received level (RL) of ship noise during instances where ships came within 50 km of a tagged beluga, first the source levels (SL) of underwater ship noise were estimated using the JOMOPANS-ECHO source level model (Table 3; MacGillivray & de Jong, 2021). SL is defined as the sound pressure level at 1 m from an ideal point source emitting the same amount of sound as the ship or other distributed source using 1 μ Pa as the reference pressure (Au, 1993). The JOMOPANS-ECHO model calculates the ship SL spectrum in decidecade bands as a function of frequency, speed, ship length, and AIS ship type. Vessel classification and length were obtained from the MarineTraffic.com web service based on the maritime mobile service identity (MMSI) number. The statistical uncertainty, reported as the standard deviation of the deviation between model and measurement in the predicted source level spectra of the model, is estimated to be 6 dB (MacGillivray & de Jong, 2021). The propagation loss of underwater noise from different classes of vessels that were transiting the area was estimated based on modeled SLs (dB re 1 μ Pa), and then applied to actual ship tracks and measured ship speeds (derived from AIS data). RL is the result of the SL minus the transmission loss (TL; Urick, 1983) such that

$$RL = SL - TL$$
(1)

Approximate transmission loss for ships is estimated assuming a combination of spherical and cylindrical spreading plus frequency-dependent attenuation caused by absorption (Au & Hastings, 2008) such that.

$$TL = 15\log_{10}(R) + \alpha R \tag{2}$$

where *R* is the distance to the ship and α is a constant defined as the frequency dependent absorption (Francois & Garrison, 1982). The geometric spreading coefficient of $15\log_{10}(R)$ was used based on previous acoustic propagation modeling in the Pacific Arctic (Halliday, Pine, et al., 2021). *R* was calculated continuously as the distance between a whale's location and a ship over the duration of the encounter event using the "Points to Line" tool in ArcMap 10.8. For encounters where more than one ship was located within 50 km of the whale, an estimated combined RL was calculated by converting received level for each individual ship to pressure, summing received pressure levels of all individual ships within 50 km, and then converting the total back to decibels. The calculated RL values apply to a receiver well below the surface, and are roughly the maximum RL in the water column; levels of ship noise (especially the lower frequency components) received by any belugas close to the surface will be reduced by pressure release effects (Urick, 1983). We did not account for the vertical location of the receiver in the water column and acknowledge there are uncertainties involved in these approximate estimates of RL; the RL estimate is intended to provide context for encounters, rather than serving as an absolute RL. We caution against using modeled values for estimating absolute RL, especially for data sets that might be used to establish thresholds for disturbance.

TABLE 3 Summary of ship metrics for ships involved in known encounters with belugas at distances less than 50 km. Source levels (broadband SL; dB at 1 m μ Pa) were modeled using the JOMOPANS-ECHO source level model using recordings of ship noise from the Port of Vancouver's ECHO program (MacGillivray & de Jong, 2021). Decidecade band source levels (dB re 1 μ Pa) were estimated in the 10 Hz-32 kHz band and ship speed was set at a value of 10 knots only for this Table's comparison.

Vessel name	MMSI	Vessel type	Length (m)	SL (dB)	Encounters (<50 km)	n Whales
Andrey Pervozvanniy	273414670	Tanker	169	175.6	1	1
Arkadiy Chernyshev	273359930	Cargo	113	171.0	2	2
BBC Oregon	305462000	Cargo	139	172.8	1	1
Boris Sokolov	209387000	Tanker	214	177.7	1	1
David Thompson	316001090	Research/Government	29	167.0	2	2
Fathom Wave	316032737	Tug	19	180.1	2	2
Frosti	316001821	Fishing	39	174.4	3	1
Georgiy Brusilov	212770000	Tanker	299	180.6	2	2
High Progress	636012730	Tanker	183	176.3	3	2
Kelly Ovayuak	316004160	Tug	45	187.6	2	1
Mirai	431939000	Research/Government	128	179.9	1	1
Nordic Olympic	356986000	Cargo	225	177.0	1	1
Rubin	273189700	Tug	58	189.8	1	1
Rudolf Samoylovich	311000627	Tanker	299	180.6	1	1
Sikuliaq	338417000	Research/Government	79	175.7	1	1
Sir Wilfrid Laurier	316052000	Research/Government	83	176.1	5	2
St. Confidence	273443870	Cargo	98	169.8	1	1
Vladimir Vize	477194200	Tanker	299	180.6	1	1

2.4 | Statistical and qualitative analyses

All whale locations where one or more ships were less than 50 km from a tagged beluga were classified as an "impact" time segment. Whale locations were grouped together in time series based on consecutive locations when a ship was within 50 km of a tagged beluga and included an equivalent or near equivalent number of whale locations both before and after the "impact" time segments based on the duration of the "impact" segment. These files were then categorized as "before," "during," and "after"; where the "before" segment was the period where the ship was approaching the 50 km radius, the "during" segment was the period where the ship was within 50 km of the tagged whale, and the "after" segment was the subsequent period where the ship was beyond the 50 km distance radius to a tagged whale.

An equal number of control time series were selected where all ships were over 125 km away from a tagged whale to control for natural variation in beluga movement behavior in the absence of vessels. These two sets of series, the ship series and the control series, allowed us to perform a pseudo before-after control-impact analysis. For the control series, whale locations were selected at equivalent times and for the same durations as the ship series (before, during, and after) for each paired encounter. A minimum of three consecutive whale locations were used to represent each before, during, and after segment for both the ship and control series. By comparing the control time series with the ship time series, natural variation in beluga movements can be compared with beluga movements during exposure to ships.

The impact of the distance of ships on beluga swim speed and bearing was analyzed using a before-after control-impact design. Beluga swim speed was estimated from the CRW modeled locations by calculating the distance (meters) between consecutive whale locations and dividing by delta time (seconds). Beluga delta bearing was estimated by calculating the difference in bearing between consecutive whale locations. We acknowledge that variable times between spatially corrected original time stamped data points could bias estimates of actual swim speed and turning angle. Linear mixed effects models in R (package: Ime4; function: Ime; Bates et al., 2015; R Core Team, 2021) were fitted separately with beluga swim speed or delta bearing as dependent variables, and the logarithm (log_{10}) of ship distance relative to the whale (meters) as a continuous independent variable. The log_{10} distance of 125,000 m was applied to control series since 125 km was the distance threshold used to identify ship presence. Ship noise received level was included as an alternate continuous independent variable and 90 dB was applied to control series, since 90 dB is roughly the lower boundary of broadband (50-1,000 Hz) ambient sound levels during the open water period (Halliday et al., 2017; Halliday, Barclay, et al., 2021; Insley et al., 2017). Ship count (number of ships within 50 km involved in the encounter), ship class and encounter exposure time (minutes) were included as additional covariates. A categorical variable was included that identified the time series as control (ship absent) or impact (ship present). Individual whale ID and a chronological encounter number were included as nested random effects. We fitted additional linear mixed effects models accounting for the interaction between ship presence/ absence and time segment (i.e., before, during, after) and removed the ship distance variable as it was correlated to the impact time segments. To account for potential temporal autocorrelation within time series, an encounter specific random effect was also included and modeled to follow a continuous time autoregressive process within each time series.

A generalized additive mixed model (GAMM) was fitted in R to identify inflection points in the relationship between ship distance and whale speed or delta bearing (package: gamm4; function: gamm; Wood & Scheipl, 2020). Residual diagnostic plots with normalized randomized quantile residuals were used to assess model fit and assumptions, and correlation tests and plots of the lagged residuals for up to 30 lags were assessed for temporal autocorrelation. Models accounting for autocorrelation were compared to those not adjusting for autocorrelation by using Akaike's information criterion (AICc) corrected for small sample sizes (package: qpcR; function: AICc; Spiess, 2014).

Beluga behavioral responses to ships were broadly grouped by visual examination of satellite location data tracks where spatial resolution permitted. Animations for each encounter event contain whale locations which were color coded by before, during, after, and control segments for ease of interpretation. Each encounter event's animation was reviewed independently by three authors (M.J.M., W.D.H., L.S.) and grouped into one of three broad behavioral response categories: potential avoidance, no lateral response and undetermined. Potential avoidance behavior was characterized as a notable change (>35°) in the direction of travel occurring in the before or during segment of a ship encounter. No lateral response was characterized as no notable change in the direction of travel occurring in the before or during segment of a ship encounter. Encounters where whale directional movement was unclear due to sparse surrounding locations or close proximity to shore (and therefore no space to swim away from the ship) were categorized as "undetermined." For each encounter, we checked that the movement trajectory was real and not an artefact of the CRW model and/or location error by removing the remaining lowest accuracy locations (Argos classes 0, A and B), refitting the CRW models and confirming that the movement trajectory associated with the CPA was reproduced.

We tested the ability to subjectively group beluga responses to encounters with ships by visual assessment of the animations using linear mixed effect models to identify relationships between behavioral response type and changes in whale bearing. Models fitted in R (package: Ime4; function: Ime; Bates et al., 2015; R Core Team, 2021) incorporated beluga delta bearing as a dependent variable, and behavioral response type as a categorical predictor variable which included "control," "potential avoidance response," "no lateral response," and "undetermined response" as factors. Individual whale ID and a chronological encounter number were included as nested random effects. To account for potential temporal autocorrelation within time series, an encounter specific random effect was also included and modeled to follow a continuous time autoregressive process within each time series. We also

assessed models that used beluga swim speed as the dependent variable. Residual diagnostic plots with normalized randomized quantile residuals were used to assess model fit and assumptions, and correlation tests and plots of the lagged residuals for up to 30 lags were assessed for temporal autocorrelation. Models accounting for autocorrelation were compared to those not adjusting for autocorrelation by using Akaike's information criterion (AICc) corrected for small sample sizes (package: qpcR; function: AICc; Spiess, 2014).

The main relationship of interest is that between behavioral response type and whale bearing. However, we accounted for the possible need to control for the time segment (i.e., before, during, after) as well as the interaction between behavioral response type and time segment. We used likelihood ratio tests (LRT) to test for the effects of the time segment and removed this variable and the interaction of behavior with time segment when there was a lack of evidence for such effects. A pairwise comparisons test using least-squares means (package: Ismeans; Lenth, 2016) was used to compare differences in whale bearing across behavioral response types.

2.5 | Beluga dive profiles

For encounter events between belugas and ships at distances less than 50 km, whale dive data were examined from the SPLASH10-F-238 tags for a 24 hr period centered around the time of CPA for each encounter. TDRs sampled depth at a 1 s frequency, and transmitted the data subsampled at a 75 s frequency. There were occasional 1 hr or greater time gaps in the dive profiles resulting from 1 hr time blocks where the TDR failed to transmit any dive data to the satellite (i.e., one message equates to one hr of continuous data). For periods with dive data, individual dives were characterized by time and depth metrics and classified into eight dive types according to the methods of Storrie et al. (2022; Table 4) which used the program divebomb (Nunes, 2019) in Python v3.7.1 (Van Rossum & Drake, 2009). Divebomb was used to measure dive parameters and categorize any dive made below a depth of 5 m. Storrie et al. (2022) defined eight dive types based on 90,211 dives reported from the 2018 tagged whales in this study (91.6% of reported dives) and seven additional EBS belugas tagged in 2019 (8.4% of reported dives). Seafloor depth for each dive was estimated from the International Bathymetric Chart of the Arctic Ocean (IBCAO) version 3.0 (Jakobsson et al., 2012) and the General Bathymetric Chart of the Oceans (GEBCO Bathymetric Compilation Group, 2020). Identified dives along with their measured parameters (i.e., dive duration, descent rate, ascent rate, bottom duration, maximum depth) were incorporated into each whale-ship encounter event and sorted by time. Dive data were not assessed quantitatively due to nonnegligible data gaps during most encounter events. Due to dive data gaps and known variability in EBS beluga dive behavior (Storrie et al., 2022), we cannot be certain whether changes in dive behavior represent disturbance responses to ships or natural behavior. The dive profiles are provided as additional context and to generate hypotheses on beluga behavior during ship encounters (see Figures S1-S18).

Storrie et al. (2022) proposed the function of classified beluga dives based on time and depth structure, foraging theory, beluga physiology, and published literature on marine mammal dive behavior. In short, EBS belugas exhibit a number of dives suggestive of foraging due to their depth, long bottom durations, and rapid descent rates, including benthic dives (Deep Benthic, DB; Intermediate Benthic, IB) and pelagic dives (Deep Pelagic V, DPV; Deep Pelagic W, DPW) (Table 4). EBS belugas frequently made Shallow V-shaped dives (SV) after deep foraging type dives, with the shallow nature and slow rates of vertical movement for this dive type supporting its use in recovery or transiting behavior (Fahlman et al., 2021; Hooker et al., 2009; Lemieux Lefebvre et al., 2018; Vacquié-Garcia et al., 2019). Other dive types were variable in form (Intermediate Pelagic, IP; Shallow W, SW) and likely represent a number of behaviors including but not limited to recovery, transiting, pelagic foraging, social behaviors, and/or navigating through ice (Table 4; Loseto et al., 2006; Martin et al., 1994; Quick et al., 2017). A final dive type, Deep Pelagic Skew (DPS), was recorded only infrequently, and may represent a pelagic foraging dive with vertical pursuit of prey or an energy efficient drift dive (Lemieux Lefebvre et al., 2018; Simon et al., 2009). Full details on dive classification, sources of error, and hypotheses on dive functions are provided in Storrie et al. (2022).



3 | RESULTS

3.1 | Summary of beluga encounters with ships

During July–December 2018, there were a total of 177 encounters between eight of the nine tagged belugas and ships within a radius of 125 km (Table 1, Figure 2). Whale LC2018#5 (referred to hereafter by whale identification number only, e.g., #5) did not encounter a ship within this distance radius; however, this animal's tag only transmitted for 17 days. The remaining eight whales' tags transmitted for longer periods (46–355 days; Table 1). These eight whales varied in their number of encounters with ships ranging from 2 to 71 (Table 1, Figure 2). Two whales (#2 and #6) had notably higher numbers of encounters with ships with 57 and 71 encounters, respectively. This was due to the near-shore migration routes traveled by those individuals (Figure 1a). Whales #2 and #6 also had the nearest CPAs to ships of all the tagged individuals examined (Table 1). During the month of November in the Chukchi Sea, the highest number of encounters with ships number of encounters (n = 47) between ships and four tagged belugas in the Amundsen Gulf and eastern Beaufort Sea (Table 1, Figure 2). July and December 2018 were the months with the lowest numbers of encounters, likely due to seasonally reduced accessibility of this region to ships.

For each tagged whale, the average CPA to a ship <125 km away was at a distance >75 km; however, there were five individuals which experienced CPAs to ships at distances ≤50 km (Table 1). When considering behavioral



FIGURE 2 Map of the 177 closest points of approach (CPA) for all encounters between tagged belugas and ships within a radius of 125 km. CPAs are color coded by month and shape coded by individual tagged beluga.

responses of belugas to ships located within 50 km, a total of 23 unique encounters occurred with one or more ships (Table 2). Of these encounters, three showed no indication of a lateral behavioral response (Figure 3a), eight displayed signs of potential avoidance responses (Figure 3b), and seven were categorized as undetermined. The remaining five encounters were considered data deficient and were not included in any statistical model due to time gaps >4 hr in whale locations occurring in the before, during, or after segments. Most encounters involved a single ship; however, six encounters contained two ships and one encounter involved three ships (Table 2). Of the 18 ships involved in these encounters, the Canadian Coast Guard research vessel *Sir Wilfrid Laurier* had the most encounters (n = 5; Table 3). There were no ships considered to be heavy icebreakers or seismic survey vessels; therefore, we did not examine whether the ships were operating in ice or if seismic ships might have been active. For the encounters within 50 km, CPAs between whales and ships were highly variable and ranged from 6.8 to 46.7 km (Table 2).

3.2 | Statistical effects

Results of the statistical models should be interpreted with caution since some variables were modeled and without experimental validation from additional controlled studies. A negative correlation was identified in the modeled variables where beluga swim speed increased with decreasing distance to ships in the encounter (i.e., impact) time series ($slope = -0.704 \pm 0.139 \text{ m/s/log}_{10}$ ship distance, $t_{966} = -5.059$, p < .001) and was different from the control time series ($t_{966} = -2.316$, p = .021; Figure 4a, Table 5), but remained relatively constant through time within the control time series ($t_{963} < 1.493$, p > .136). A perceived nonlinear change in the correlation between beluga swim speed with distance to ships was identified (GAMM: edf = 3.208, F = 22.81, p < .001). An inflection point occurred around a distance of $log_{10}4.9 \text{ m}$ (Figure 4b), which equates to a distance of $\sim779 \text{ km}$. The average swim speed was $1.92 \text{ m/s} \pm 0.34 \text{ SD}$) at $\sim13 \text{ km}$ ($log_{10}4.1 \text{ m}$) distance to ships compared to $1.07 \text{ m/s} \pm 0.70 \text{ SD}$) at $\sim79 \text{ km}$ distance to ships (Figure 4b). The second set of models examining beluga swim speed by encounter segment showed evidence that swim speed increased in the during segment ($t_{965} = 2.051$, p = .041, $M \pm SD = 1.26 \pm 0.59 \text{ m/s}$), but did not differ from the control ($M \pm SD = 1.07 \pm 0.51 \text{ m/s}$) in the before segment ($t_{965} = -0.844$, p = .399, $M \pm SD = 0.99 \pm 0.58 \text{ m/s}$) and after segment ($t_{965} = -0.191$, p = .849, $M \pm SD = 0.99 \pm 0.58 \text{ m/s}$; Figure 5; Table 5).

There was evidence that change in beluga bearing increased in encounters visually assigned as "potential avoidance response" compared to the control time series ($t_{965} = 2.800$, p = .005). There was no evidence of a difference in beluga delta bearing in encounters visually assigned as "no lateral response" compared to the control time series ($t_{965} = -0.212$, p = .833; Table 5). There was no evidence of a correlation between beluga swim speed and the visually assigned behavior type categories. There was some indication of temporal autocorrelation when analyzing residuals, and the lowest AICc values were obtained by including the continuous time autoregressive process; hence, all final models included the random effect for temporal autocorrelation.

3.3 | Encounters with a potential avoidance behavioral response

Of the eight encounters between belugas and ships where a potential avoidance behavioral response was observed, four encounters involved a single ship and four encounters included two or three ships (Table 2). The CPAs between the whales and ships ranged from 12.6 to 43.1 km and at CPA the estimated RLs of ship noise near the whale were 98–133 dB re 1 μ Pa (Table 2). A brief summary of each encounter is provided in the following paragraphs. All times are reported in UTC. In these eight cases, there were indications of two broad types of dive disturbance response: (1) shallow diving/swimming behavior from 0 to 20 m depth, and (2) anomalous spike-shaped dives, where maximum depth was reached and the beluga either immediately started to ascend or had a much shorter bottom phase than was typical for the given dive type based on measured dive parameters in Storrie et al. (2022; Table 4, Figures S1–S8).



avoidance behavioral response. For each encounter, the top panel is a bird's eye view of ship AIS and whale tracks; the closest points of approach (CPA) identified with black FIGURE 3 (a) Beluga encounter with a ship where no lateral behavioral response was observed. (b) Beluga encounter with three ships with evidence of a potential whale shows the location of the encounters as a black dot in the Pacific Arctic. The bottom two panels show the whale's dive profile with sea floor depth during the encounter and the control period. Light gray bars are placed along the top of the dive profiles for time periods when dive data were missing. A green box is placed over the "during" period locations: gray ("control" segment), orange ("before" segment), green ("during" segment), pink ("after" segment), and black (all other locations). The inset map in gray scale circles. Whale locations estimated from the CRW model are shown as color coded dots pertaining to the encounter segments and control segments. Color coded whale which includes the CPA.



FIGURE 4 (a) Scatterplot of beluga swim speed by distance to ships. A negative correlation was identified in the modelled variables indicating an increase in beluga swim speed with increasing proximity to ships in the encounter (i.e., impact) time series and was different from the control time series. (b) Generalized additive mixed model result used to identify a nonlinear change in the relationship between beluga swim speed with distance to ships. An inflection point occurred around a ship distance of log 10 4.9 m, which equates to a distance of ~79 km.

3LE 5 Linear mixed effect model results for the comparison of tagged beluga swim speed and change in bearing between the control (ship absent) and distance to shi	966), encounter time segment ($df = 965$), and behavioral response type ($df = 965$). Eighteen encounters between tagged belugas and ships consisting of 986 whale	ons were included in these models. Effect sizes are reported with 95% confidence intervals (Cl) and values based on $p < .05$ are shown in bold.
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TABLE 5 Linu $(df = 966)$, encoulocations were ir	ear mixed efft unter time se soluded in the	ect model results f gment ($df = 9.65$), sse models. Effect	for the co and beha sizes are	mparison of tagged b vioral response type (reported with 95% cc	eluga swim sl df = 965). Ei infidence int	peed and char ghteen encour ervals (Cl) and	ige in beari iters betwe values bas	ng between the contr sen tagged belugas an ed on <i>p</i> < .05 are shov	ol (ship abser d ships consi vn in bold.	it) and distance sting of 986 wha	to ships ale
Swim speed by	distance to sh	hips		Swim speed by encou	unter segmen	t		Delta bearing by resp	onse type		
Predictors	Estimates	G	d	Predictors	Estimates	Ū	d	Predictors	Estimates	C	d
(Intercept) [Control]	4.6	3.21, 5.98	<.001	(Intercept) [Control]	1.03	0.85, 1.21	<0.001	(Intercept) [Control]	16.52	10.33, 22.71	<.001
Ship present	-0.15	-0.28, -0.02	0.021	Segment: Before	-0.06	-0.19, 0.08	0.399	Behavior: Avoidance	6.75	2.02, 11.48	0.005
log ₁₀ distance	-0.7	-0.98, -0.43	<.001	During	0.15	0.01, 0.30	0.041	No lateral response	-1.04	-10.68, 8.60	0.833
				After	-0.02	-0.17, 0.14	0.849	Undetermined	6.07	-0.07, 12.20	0.053



FIGURE 5 Beluga changes in swim speed (meters/second) estimated at the times of the CRW-modeled locations during each of the encounter segments "before," "during," and "after" based on when one or more ships were within 50 km of a tagged whale, as well as during control segments when no ships were present within a 125 km radius. Violin plots are used to show the probability density of the data at different values. Boxes inside each violin plot represent the interquartile range, the line within the boxes is the median, whiskers are the minima and maxima, and dots are outliers.

3.3.1 | Potential avoidance response 1

On August 7, 2018, whale #2 was in the Amundsen Gulf and traveling in a westward direction (heading ~260°) toward shore (Figure S1). This was the first known encounter between whale #2 and a vessel at \leq 50 km distance during the study period. Two ships, a fishing/government research vessel, *Frosti*, and a coast guard/government research vessel, *Sir Wilfrid Laurier* (Tables 2 and 3), were traveling in a southeasterly direction in tandem (see Animation S1). The CPA between the whale and the two ships occurred at 22:27 UTC at a distance of 19.1 km (based on *Frosti's* position) with an estimated combined ship noise broadband RL near the whale of 98 dB re 1 µPa (Table 2). A U-shaped diversion in lateral movement in the during segment added a distance of 12 km to the whale's westward transit to shore. During the hour prior to the CPA, the beluga made three anomalous spike-shaped dives with short bottom durations (Figure S1).

3.3.2 | Potential avoidance response 2

On August 20, 2018, whale #2 was located in the Amundsen Gulf when it encountered the tugboat *Kelly Ovayuak* (Tables 2 and 3). The CPA between the whale and the ship occurred at 14:58 UTC at a distance of 12.6 km, and the estimated received level from the ship near the whale was 127 dB re 1 μ Pa (Table 2). There were no dive data from 14:00 to 19:00 UTC, including during the CPA. Prior to the CPA, the dive profile indicated foraging behavior with alternating DB and SV dives. The whale appeared to be foraging, and then exhibited a directed lateral movement north and away from the ship track in the during segment (see Animation S2, Figure S2).

3.3.3 | Potential avoidance response 3

On August 25, 2018, whale #2 was traveling in a northwesterly direction (heading ~320°) in the Amundsen Gulf (Figure 3b). Three ships, *Frosti* (fishing/government research vessel), *High Progress* (tanker ship), and *Sir Wilfrid Laurier*

(coast guard/government research vessel) (Tables 2 and 3), were traveling in an easterly direction overall; however, the *Frosti* changed heading several times likely due to fishing activity (see Animation S3). The CPA between the whale and the three ships occurred at 03:48 UTC at a distance of 13.1 km (based on *Frosti*'s position) with an estimated combined ship noise broadband RL of 119 dB re 1 μ Pa (Table 2). During the hour prior to the CPA, the beluga made seven consecutive SV dives to 6.5–20.5 m depth. At 04:02 the animal returned to the surface and did not descend below 3 m depth for the next full hour (Figure 3b). These shallow dives/subsurface swimming coincided with a distinct westward change in lateral movement (heading changed from ~320° to 280°) in the during segment (Figure 3b).

3.3.4 | Potential avoidance response 4

On November 8, 2018, whale #4 was located in the Chukchi Sea near the east coast of Russia in shallow water (~50 m depth; Figure S4). A cargo ship, the *Arkadiy Chernyshev* (Tables 2 and 3), was traveling southeast parallel to the coast (see Animation S4). The CPA between the whale and the ship occurred at 07:27 UTC at a distance of 40.8 km, and the estimated broadband RL from the ship was 105 dB re 1 μ Pa near the whale (Table 2). Half an hour prior to the CPA, the whale made two shallow dives (SW and SV) followed by an anomalous spike-shaped IB dive to 42 m depth immediately prior to the CPA at 07:17 UTC, which lasted for 8.5 min but had no bottom duration (i.e., zero time spent at the dive's maximum depth). These dives coincided with an eastward change in the whale's lateral movement (heading ~50°), toward open water and away from the ship's track in the during segment (Figure. S4).

3.3.5 | Potential avoidance response 5

On August 12, 2018, whale #6 was located in the Amundsen Gulf moving southwest in ~350 m deep water (Figure S5). This was the first known encounter between whale #6 and a vessel at \leq 50 km distance during the study period. A coast guard/government research vessel, *Sir Wilfrid Laurier*, was heading in an eastward direction in tandem behind the track of a tugboat, *Fathom Wave*, trailing by approximately 400 m (Tables 2 and 3; see Animation S5). The CPA (based on *Fathom Wave*'s position) occurred at 06:32 UTC at a distance of 43.1 km. The estimated combined RL near the whale from both ships during the CPA was 99 dB re 1 µPa (Table 2). In the during segment, the whale did not alter its dive behavior and performed a series of DB and DPV dives to 400–500 m depth (Figure S5). There was a distinct change in lateral movement where the whale appears to have moved in a northeastward direction in the before segment, followed by a southwestward directed movement in the during segment, creating a circular shaped movement path.

3.3.6 | Potential avoidance response 6

This encounter occurred on August 29, 2018. Whale #6 was westbound in the Beaufort Sea when it encountered the Canadian coast guard/government research vessel *David Thompson* (Tables 2 and 3). The CPA between the whale and the ship occurred at 09:15 UTC at a distance of 20.9 km, and the estimated received level from the ship near the whale was 100 dB re 1 μ Pa (Table 2). There were no dive data from 05:00 to 09:00 or from 10:00 to 19:00 UTC, with only 1 hr of dive data during the CPA (09:00 to 10:00). The whale made one DB dive until 9:15 UTC and then remained at the surface with no additional dives for the next 45 min (Figure S6). In the during segment, the whale appeared to backtrack along its previous route and then moved in a southward direction away from the ship track (see Animation S6). In the after segment the whale returned to its previous westbound trajectory. The changes in lateral movement in the during and after segments created a circular shape in the whale's movement path (Figure S6).

3.3.7 | Potential avoidance response 7

On September 12, 2018, whale #6 was located in deep water (~700 m depth) in the western Alaskan-Beaufort Sea moving in a northward direction. A research vessel, *Sikuliaq* was traveling in a southeasterly direction and a tanker ship, *High Progress* (Tables 2 and 3), was headed in a southwesterly direction. The CPA between the whale and the closest ship (*Sikuliaq*) occurred at 08:44 UTC at a distance of 14.3 km, and the combined estimated RL from the ships, for a location near the whale, was 112 dB re 1 μ Pa (Table 2). Between 01:00 and 07:00 UTC, whale #6 was traveling in a northward direction. At 07:30, the whale made two anomalous 'spike' shaped dives which included a DPWS dive to 163 m depth with no bottom duration directly followed by a DB dive to 1,010 m depth with a short bottom duration of 3.75 min. The whale then remained at the surface and did not descend below 4 m depth from 07:53 to 08:08 which coincided with the start of a distinct westward change in the animal's heading (from ~320° to 280°) away from the ship (Figure S7). The whale remained at the surface from 08:17 to 08:53 and did not descend below 4 m depth during this time. These shallow dives/subsurface swimming coincided with the westward change in lateral movement in the during segment (see Animation S7).

3.3.8 | Potential avoidance response 8

On November 19, 2018, a tanker ship, *Georgiy Brusilov* (Tables 2 and 3), was moving northwest parallel to the Russian coastline approximately 40 km offshore in the Chukchi Sea. The CPA occurred at 06:16 UTC at a distance of 13.4 km from whale #6 (Table 2). During the CPA, the estimated RL from the ship near the whale was 133 dB re 1 μ Pa. There are gaps in the dive data from 05:00 to 06:00 and from 07:00 to 12:00 UTC; however, all dives identified over 24 hr centered around the CPA were IB dives to the seafloor (~48 m depth) indicating foraging behavior (Figure S8). There was a distinct change in lateral movement in the before segment where the whale began swimming in the opposite direction (see Animation S8). There was another distinct change in lateral movement in the during segment where the whale moved in a northeastward direction away from the ship track. The changes in lateral movement in the before and during segments created a circular shape in the whale's movement path (Figure S8).

3.4 | Encounters with no lateral response (NLR)

There were three encounters between belugas and ships when we identified no apparent lateral behavioral responses (NLR), defined as no clear change in lateral movement. All three of these encounters involved a single ship (Table 2). The CPAs between the whales and ships ranged from 23.6 to 46.7 km and estimated RLs near the whale ranged from 104 to 113 dB re 1 μ Pa (Table 2). Foraging dive types (DB, IB) were recorded in the encounter segments; however, there are nonnegligible gaps in the dive data. A summary of each encounter is provided in the following paragraphs.

3.4.1 | No lateral response 1

On August 18, 2018, whale #2 was located in Amundsen Gulf headed in a southeasterly direction (heading ~120°; Figure 3a). A research vessel, *Sir Wilfrid Laurier* (Tables 2 and 3), was headed in a southwesterly direction. The AIS ship locations were intermittent around the time of the CPA which occurred at 02:34 UTC at a distance of 23.6 km (Table 2). The estimated RL from the ship near the whale during the CPA was 113 dB re 1 μ Pa. The whale appeared to be foraging and the dive profile did not change prior to, or during, a period of two hr around the CPA (Figure 3a). Dives were long (~17.5 min) DB dives to the seafloor (~475 m depth) including dives at 02:13 and 02:37 UTC. There was no apparent change in the lateral movement path of the whale during the ship encounter (see Animation S9).

3.4.2 | No lateral response 2

This encounter occurred on November 8, 2018. Whale #2 was located in the Chukchi Sea moving southeast in parallel with the cargo ship *Arkadiy Chernyshev* (Tables 2 and 3). The whale was located between the ship and land (Figure S10). The CPA between the whale and the ship occurred at 04:56 UTC at a distance of 46.7 km, and the estimated received level from the ship near the whale was 104 dB re 1 μ Pa (Table 2). Post ship encounter, the whale turned to an easterly direction toward the ship track and open water (see Animation S10). In the during segment of this encounter, there was only 1 hr of dive data (05:00–06:00) which contained four IB dives and one SW dive (Figure S10).

3.4.3 | No lateral response 3

This encounter occurred on August 24, 2018. Whale #6 was located in the Beaufort Sea when it encountered the tanker ship *High Progress* (Tables 2 and 3; Figure S11). The CPA between the whale and the ship occurred at 04:33 UTC at a distance of 39.3 km, and the estimated received level from the ship near the whale was 112 dB re 1 μ Pa (Table 2). There were no dive data from 02:00 to 03:00 and 04:00 to 06:00 UTC, including during the CPA. When dive data were available, other time periods indicated foraging with only DB and SV recorded dives. Throughout the encounter, the whale was moving in a southwesterly direction away from the ship track with no directional change in lateral movement (see Animation S11).

3.5 | Encounters with undetermined behavior (UND)

There were seven encounters between belugas and ships in which the behavioral responses were considered to be undetermined due to unclear whale lateral movements through time. For an encounter to be categorized as undetermined, whale locations were in close proximity to shore or whale points were sparse outside of the control and impact segments to a degree where a change in lateral behavior could not be visually assessed with confidence (see Animations S12–S18, Figures S12–S18). Whales #2 and #4 were involved in undetermined encounters (Table 2). Four encounters involved a single ship and three encounters included two ships. The CPAs between the whales and ships ranged from 6.8 to 39.4 km and estimated RLs near the whales ranged from 108 to 159 dB re 1 µPa (Table 2).

3.6 | Encounters determined to be data deficient (DD)

There were five encounters between belugas and ships that were considered to be data deficient due to significant gaps (>4 hr) in whale location data during the encounter segments. Four whales (#1, 2, 6, and 8) were involved in data deficient encounters (Table 2). The encounters each involved a single ship. The CPAs between the whales and ships ranged from 25.2 to 45.8 km and estimated RLs at a location near the whales ranged from 97 to 126 dB re 1 μ Pa (Table 2).

4 | DISCUSSION

This study summarizes the number of instances (n = 177) ships were encountered by tagged belugas and their behavioral responses during the period July–December 2018 in the Pacific Arctic (Table 1). We provide correlational evidence that belugas showed behavioral responses to vessels in the Pacific Arctic based on an increase in swim speed and change in bearing when in variable range of one or more ships. Results of the linear mixed effects models on modeled variables provide evidence that beluga swim speed was faster in the during segment of encounters (i.e., when ships \leq 50 km from the whale) compared to the control, before, and after segments (Table 5, Figure 5). A correlation between an increase in beluga swim speed in the presence of ships was estimated to occur up to ~79 km distance (Figure 4). Additional model results provide evidence that change in beluga bearing increased in encounters visually assigned as "potential avoidance response" (n = 8 encounters) compared to the control time series. Furthermore, there was no evidence of a difference in beluga delta bearing in encounters visually assigned as "no lateral response" (n = 3 encounters) compared to the control time series.

A flee or avoidance response in belugas has been documented repeatedly at distances >10 km from ships (Finley et al., 1990; Miller et al., 2005; this study), suggesting that the response was to noise given that these distances are beyond the whales' visual and echolocation detection ranges. Finley et al. (1990) reported that belugas altered their acoustic behavior and began producing alarm calls when a transiting icebreaker vessel was approaching at 80 km distance. The whales further responded by fleeing when the icebreakers were at distances of 35–50 km with broadband RLs ranging from 94 to 105 dB re 1 μ Pa (Finley et al., 1990). Erbe and Farmer (2000) modeled the zones of acoustic impact around icebreaker ships affecting Arctic belugas and found that zones of disturbance were only slightly smaller than predicted zones of audibility (35–78 km distance, depending on location), which supports earlier conclusions by Cosens and Dueck (1993) and Richardson et al. (1995). Further, based on a propagation model Schack and Haapaniemi (2017) estimated that belugas could potentially detect ship noise from container and icebreaker vessels up to distances of 48 km and 57 km, respectively, during the ice-covered season and up to 75 km and 79 km, respectively, in open water. Results of the statistical analysis between beluga swim speed and ship distance corroborate these findings, with a negative correlation between increasing swim speed with decreasing ship distance up to approximately 79 km (Figure 4).

In this study, potential avoidance responses were observed in eight encounters at varying distances (12.6-43.1 km) with estimated maximum RLs of 98–133 dB re 1 μ Pa from a variety of ship types and sizes (Tables 2 and 3). This again raises the question as to whether received level, signal to noise ratio, signal type, or a combination of these elicit a flee response in belugas in the Arctic (e.g., Richardson et al., 1995). The wide range of estimated maximum received levels also calls into question the applicability of a single noise threshold, such as the 120 dB re 1 μ Pa disturbance threshold established by the National Oceanic and Atmospheric Administration (NOAA; National Marine Fisheries Service, 2018), since the tagged belugas in this study likely reacted at much lower received levels in agreement with the findings of previous studies. However, it should be noted that the received level estimates in this study include a degree of error, given that the source levels used were modeled based on different ships measured in other areas (i.e., the Port of Vancouver's ECHO program) and the propagation loss calculations were relatively simplistic and did not include water depth or bottom conditions along the acoustic propagation path, or the location of the beluga in the water column. It is important to note; however, that the relative differences in received levels is still useful for comparisons between different distances and vessel types, but caution should be used when examining the absolute values.

Two belugas (whale #2 and #6) were involved in seven of the eight encounters exhibiting potential avoidance behavior (Table 2). Comparatively, these two whales were present in the three encounters where no lateral behavioral response was found (Table 2). Both whale #2 and #6 were consequently represented frequently in case studies due to their higher number of encounters with ships during the study period (Table 1). Currently, it is unknown if there is an individual or group-related noise threshold where some belugas exhibit an avoidance response whereas others might have no reaction.

It is possible that whales became more tolerant to ship noise as the season progressed, or that costs associated with disrupted foraging outweigh the costs of avoiding a perceived threat in certain encounters. There is potential for some level of habituation or desensitization to ships for whales with repeated encounters; however, this is difficult to assess in the current study due to possible confounding factors, low sample sizes, and because tag spatial resolution declined toward the end of the study period. Moreover, it is important to consider that all tagged individuals were adult male belugas. These individuals likely encountered ships in previous years in the region, and the potential level of seasonal or permanent habituation to ship noise is unknown.

Additional observations support that belugas are more tolerant of stationary, constant noise sources compared to dynamic noise sources such as an approaching vessel (Fraker, 1977, 1978; McCarty, 1981; Stewart et al., 1982, 1983). Fraker (1978) and Stewart et al. (1983) also report instances where feeding belugas did not react to approaching ships or underwater playbacks of drilling sounds, respectively. The encounters where no lateral behavioral response was observed were chronologically intermixed with encounters from the same individuals where a potential avoidance response was observed (Table 2). CPA distance between the whales and ships and the estimated RLs in nonresponsive encounters were similar to encounters where potential avoidance responses were identified (Table 2). In three encounters with a potential avoidance response, the whales exhibited diving behavior characteristic of foraging prior to ship approach, which appeared to be disrupted towards the CPA (Figures 3b, S4, S7). Our findings corroborate previous studies which report disrupted foraging behavior in the presence of ships or ship noise for additional cetacean species (e.g., New et al., 2020; Pirotta et al., 2015; Steckenreuter et al., 2011; Wisniewska et al., 2018). However, exhibiting different reactions to similar noise levels suggests context-dependent responses in this species. Gomez et al. (2016), Richardson et al. (1995), and Southall et al. (2007) found that noise level failed to reliably predict identifiable behavioral responses in some marine mammals, as responses were affected by the context of the exposure and by the animal's experience with acoustic disturbances and motivation.

In six out of eight encounters that indicated a potential avoidance response, belugas demonstrated some degree of shallow diving behavior (SV, SW, and some IP type dives) or subsurface swimming, representative of transiting behavior in the during segment (Storrie et al., 2022). When paired with the lateral movement responses, this may strengthen the evidence of a flee response from ships, as observed in previous studies (Finley et al., 1990; Miller et al., 2005).

A second possible indicator of a dive disturbance response were the anomalous spike-shaped dives which occurred in the during segment in three of the potential avoidance encounters (Figures S1, S4, S7). Maximum depths between 20 and 1,010 m were reached, and the beluga either immediately started to ascend or had a much shorter bottom phase than expected for the given dive type (Table 4; Storrie et al., 2022). Belugas in the Arctic and sub-Arctic previously have been observed to perform long duration dives as an avoidance response to ships (Blevins, 2015; Finley et al., 1990; Krasnova et al., 2009); however, the present study provides the first evidence of a potential deep dive response down to 1,010 m depth (Figure S7). Similarly, northern bottlenose whales (Hyperoodon ampullatus) were reported to undertake nonforaging dives (evidenced by lack of foraging echolocation clicks) to 2,339 m depth in response to naval sonar (Miller et al., 2015). Such deep dives could represent an initial cryptic escape response thought to have evolved to reduce detection by predatory killer whales (Orcinus orca, Miller et al., 2015). Killer whales are not commonly found in the eastern Beaufort Sea and Amundsen Gulf where belugas were during July-September of this study, and are not considered to have elicited this dive response, at least in the earlier encounters of this study (Higdon et al., 2013; Stafford, 2019). Alternatively, the relatively short bottom phases of the deep spike dives could represent abandoned foraging behavior similar to that exhibited in beaked whales exposed to naval sonar and recordings of killer whale vocalizations (Tyack et al., 2011). Diving to several hundred meters as an escape response and disruption of likely foraging activity may be energetically costly to belugas, but only over relatively short periods with unknown long-term consequences.

The dive types characterized and classified by their time- and depth-structures in Storrie et al. (2022) exhibit within-group variability, and occur through the annual cycle often in regions which currently experience little to no anthropogenic activity. Hence, it is likely that certain dive types observed during ship encounters represent several behaviors rather than solely a response to vessels. Future studies will require the incorporation of ancillary data on animal acoustics, orientation, and/or acceleration to enable identification of foraging behavior within dives (e.g., Miller et al., 2015; Tyack et al., 2011) to confirm whether these deep dives represent a cryptic escape response, disrupted foraging behavior, a natural part of beluga behavior, or some combination.

Changes in a cetacean's lateral and vertical movements can be caused by other factors including social cues from conspecifics. Another social odontocete, the sperm whale (*Physeter macrocephalus*), has been observed making sudden turns during directed movements, which Whitehead (2016) suggests could be in response to receiving

information on foraging success from another member of the group. The possibility that belugas exhibit the same behavior cannot be discarded; however, the timing of the movements described herein relative to ships and the consistency of beluga foraging behavior within a season (Storrie et al., 2022) indicates that avoidance is a more parsimonious explanation. Furthermore, EBS belugas are harvested by Inuvialuit in estuaries and near-shore areas during July and August (Harwood et al., 2014a). The seasonal whaling camps clustered on the coast of the Mackenzie River delta and Paulatuk when a change in lateral movement was observed were >100 km away, so responses in August 2018 were more likely due to the larger vessels analyzed herein than Inuvialuit harvesters.

Repeated anthropogenic disturbance likely has led to changes in local distributions of beluga populations (see Blevins, 2015 for a review). Therefore, increased shipping activity in key regions also may result in shifted distributions, at least in some areas. During migration along the coast of Alaska, belugas have shown a greater degree of displacement where there is extensive active subsistence hunting by local Indigenous communities (Burns & Seaman, 1986; Huntington, 2000; Stanek, 1996), which suggests avoidance behavior associated with experience. However, this may not be the case in other areas such as the Mackenzie River estuary, where thousands of belugas migrate each summer (Harwood et al., 2014b) despite local subsistence hunting each year (Fisheries Joint Management Committee, 2013). Consequences of anthropogenic disturbance that do not cause physical harm, such as belugas avoiding an area, may seem inconsequential; however, displacement from important habitats including feeding or calving grounds could be harmful to the sustainability of this species (Hobbs et al., 2006). For encounters where an avoidance behavioral response was detected, three encounters (A3, A4, A7; Figures S3, S4, S7) indicate that the whales' perceived foraging behavior ceased for a period of one to two hours during or when approaching the CPA. This could represent a 4.2%-8.3% reduction in available time spent foraging on a given day if the animal reacted to the ship stimuli. Such values are an estimate but provide insight to an amount of lost foraging effort elicited by single disturbance events in this study and provide context for future increases in ship traffic and expected encounter rates.

There are certain caveats associated with this study that need to be considered. These include (1) the inherent spatial accuracy in some whale location data derived from the Argos satellite system and time between original time stamped data points. While the crawl model accounts to some degree for issues associated with spatial accuracy, variable times between spatially corrected original time stamped data points could bias estimates of actual swim speed and turning angle. Five encounters between belugas and ships were classified as data deficient and could not be investigated further primarily due to insufficient tag data (Table 2). (2) The received level estimates were also approximations and include a degree of inaccuracy due to data not being available on the source level of each ship during the encounter, the simplistic propagation modeling used, and variable depths of the tagged whales in the water column. Future work would benefit from additional focused recordings of underwater ship noise unique to each vessel, including measurements of source levels. To accurately assess the received level for each individual whale would require the use of acoustic tags or an extensive array of acoustic recorders, which would not be feasible with currently available technology for the long-term tagging period of this study and the wide geographic scope of the encounters with ships. With developments in tag technology, future incorporation of 3-dimensional movement and acoustic data streams would provide the opportunity to examine changes in the acoustic behavior of belugas, identify the acoustic signature, received level and exact time when ship noise is received at the whale, and ultimately allow a more in-depth examination of belugas' behavioral response to the type and received level of ship noise and other sounds. (3) It is possible that there were ships present which did not carry AIS transponders. For example, Halliday et al. (2018) found that only 32% of pleasure craft (i.e., private yachts, sailboats) and 70% of passenger ships traveling in the Inuvialuit Settlement Region (western Canadian Arctic) during 2012-2015 were broadcasting AIS signals. Most small local boats and many tugs operating in this region similarly do not carry AIS transponders (Halliday et al., 2020b). We may therefore have underestimated the total number of beluga encounters with ships within the 125 km radius, and as such our counts represent minimum estimates. Individual encounters could also include additional ships that were not accounted for in the AIS data set.

Finley et al. (1990) reported that responses to ship noise by both belugas and narwhals (*Monodon monoceros*) at long ranges up to 80 km may be explained in part by the fact that no similar field studies previously were conducted in pristine marine environments with industrially naive populations of marine mammals. We provide evidence that belugas in the Pacific Arctic are still reacting to ship noise at long ranges and low received levels despite a doubling or tripling of ship traffic over the past three decades (Dawson et al., 2018) and high levels of oil and gas activity in the Beaufort and Chukchi Seas during the same period (Reeves et al., 2012). Our findings corroborate previous studies showing that belugas in the Arctic often react to ships far beyond the whales' visual range, implying that the whales are reacting to the ships' underwater sound stimuli. Richardson et al. (1990) hypothesized that reaction distances of belugas will be larger when anthropogenic noise contains higher frequency (>1 kHz) components due to their sensitive high-frequency hearing. Cosens and Dueck (1993) confirmed the presence of higher frequency (5 kHz band) components in the noise signal from the icebreaker ship studied by Finley et al. (1990), and Erbe and Farmer (2000) and Schack and Haapaniemi (2017) provided further evidence that belugas should be able to detect such sounds at large distances (35–78 km and 43–79 km, respectively).

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REFERENCES

- Aguilar Soto, N., Johnson, M., Madsen, P. T., Tyack, P. L., Bocconcelli, A., & Fabrizio Borsani, J. (2006). Does intense ship noise disrupt foraging in deep-diving Cuvier's beaked whales (*Ziphius cavirostris*)? *Marine Mammal Science*, 22(3), 690– 699. https://doi.org/10.1111/j.1748-7692.2006.00044.x
- Arce, F., Bestley, S., Hindell, M. A., McMahon, C. R., & Wotherspoon, S. (2019). A quantitative, hierarchical approach for detecting drift dives and tracking buoyancy changes in southern elephant seals. *Scientific Reports*, 9(1), Article 8936. https://doi.org/10.1038/s41598-019-44970-1
- Arveson, P. T., & Vendittis, D. J. (2000). Radiated noise characteristics of a modern cargo ship. Journal of the Acoustical Society of America, 107(1), 118–129. https://doi.org/10.1121/1.428344
- Au, W. (1993). The sonar of dolphins. Springer-Verlag.
- Au, W. W. L., & Hastings, M. C. (2008). Principles of marine bioacoustics. Springer.
- Aulanier, F., Simard, Y., Roy, N., Gervaise, C., & Bandet, M. (2017). Effects of shipping on marine acoustic habitats in Canadian Arctic estimated via probabilistic modeling and mapping. *Marine Pollution Bulletin*, 125(1–2), 115–131. https:// doi.org/10.1016/j.marpolbul.2017.08.002
- Awbrey, F. T., Thomas, J. A., & Kastelein, R. A. (1988). Low frequency underwater hearing sensitivity in belugas, Delphinapterus leucas. Journal of the Acoustical Society of America, 84(6), 2273–2275. https://doi.org/10.1121/1.397022
- Bates, D., Mächler, M., Bolker, B., & Walker, S. (2015). Fitting linear mixed-effects models using lme4. Journal of Statistical Software, 67(1), 1–48. https://doi.org/10.18637/jss.v067.i01
- Bennett, M. M., Stephenson, S. R., Yang, K., Bravo, M. T., & De Jonghe, B. (2020). The opening of the Transpolar Sea Route: Logistical, geopolitical, environmental, and socioeconomic impacts. *Marine Policy*, 121, Article 104178. https://doi.org/ 10.1016/j.marpol.2020.104178
- Blevins, R. E. (2015). Sound and human impacts on beluga whales in Cook Inlet, Alaska [Doctoral dissertation]. University of Alaska Fairbanks.
- Burns, J. J., & Seaman, G. A. (1986). Investigations of belukha whales in coastal waters of western and northern Alaska. II. Biology and ecology. OCSEAP Final Report, 56(1988), 221–357.
- Castellini, M. A., Davis, R. W., & Kooyman, G. L. (1988). Blood chemistry regulation during repetitive diving in Weddell seals. *Physiological Zoology*, 61(5), 379–386. https://doi.org/10.1086/physzool.61.5.30161259
- Castellote, M., Mooney, T. A., Quakenbush, L., Hobbs, R., Goertz, C., & Gaglione, E. (2014). Baseline hearing abilities and variability in wild beluga whales (Delphinapterus leucas). Journal of Experimental Biology, 217(Pt 10), 1682–1691. https:// doi.org/10.1242/jeb.093252
- Chou, E., Southall, B. L., Robards, M., & Rosenbaum, H. C. (2021). International policy, recommendations, actions and mitigation efforts of anthropogenic underwater noise. Ocean & Coastal Management, 202, Article 105427. https://doi.org/ 10.1016/j.ocecoaman.2020.105427

- Citta, J. J., Lowry, L. F., Quakenbush, L. T., Kelly, B. P., Fischbach, A. S., London, J. M., Jay, C. V., Frost, K. J., Crowe, G. O. C., Crawford, J. A., & Boveng, P. L. (2018). A multi-species synthesis of satellite telemetry data in the Pacific Arctic (1987– 2015): Overlap of marine mammal distributions and core use areas. *Deep Sea Research Part II: Topical Studies in Oceanography*, 152, 132–153. https://doi.org/10.1016/j.dsr2.2018.02.006
- Citta, J. J., Okkonen, S. R., Suydam, R. S., Quakenbush, L., Bryan, A. L., & Olnes, J. (2020). Beluga dive behavior relative to fronts and stratified layers near Barrow Canyon, Alaska. Deep Sea Research Part I: Oceanographic Research Papers, 165, Article 103392. https://doi.org/10.1016/j.dsr.2020.103392
- Clark, C. W., Ellison, W. T., Southall, B. L., Hatch, L., Van, S., Parijs, A. F., & Ponirakis, D. (2009). Acoustic masking in marine ecosystems as a function of anthropogenic sound sources. Report SC/61/E10 presented to the Scientific Committee of the International Whaling Commission.
- Collecte Localisation Satellites. (2016). Argos user's manual. https://www.argos-system.org/
- Cosens, S. E., & Dueck, L. P. (1988). Responses of migrating narwhal and beluga to icebreaker traffic at the Admiralty Inlet ice-edge, NWT in 1986. In W. M. Sackinger & M. O. Jeffries (Eds.), Port and ocean engineering under Arctic conditions (pp. 39–54). Geophysical Institute, University of Alaska.
- Cosens, S. E., & Dueck, L. P. (1993). Icebreaker noise in Lancaster Sound, NWT, Canada: Implications for marine mammal behavior. Marine Mammal Science, 9(3), 285–300. https://doi.org/10.1111/j.1748-7692.1993.tb00456.x
- Council of Canadian Academies. (2019). Canada's Top Climate Change Risks. The Expert Panel on Climate Change Risks and Adaptation Potential.
- Dawson, J., Pizzolato, L., Howell, S. E., Copland, L., & Johnston, M. E. (2018). Temporal and spatial patterns of ship traffic in the Canadian Arctic from 1990 to 2015. Arctic, 71(1), 15–26. https://doi.org/10.14430/arctic4698
- Diachok, O. I. (1976). Effects of sea ice ridges on sound propagation in the Arctic Ocean. Journal of the Acoustical Society of America, 59(5), 1110–1120. https://doi.org/10.1121/1.380965
- Doniol-Valcroze, T., Lesage, V., Giard, J., & Michaud, R. (2011). Optimal foraging theory predicts diving and feeding strategies of the largest marine predator. *Behavioral Ecology*, 22(4), 880–888. https://doi.org/10.1093/beheco/arr038
- Duarte, C. M., Chapuis, L., Collin, S. P., Costa, D. P., Devassy, R. P., Eguiluz, V. M., Erbe, C., Gordon, T. A. C., Halpern, B. S., Harding, H. R., Havlik, M. N., Meekan, M., Merchant, N. D., Miksis-Olds, J. L., Parsons, M., Predragovic, M., Radford, A. N., Radford, C. A., Simpson, S. D., ... Juanes, F. (2021). The soundscape of the Anthropocene ocean. *Science*, 371(6529), Article eaba4658. https://doi.org/10.1126/science.aba4658
- Dujon, A. M., Lindstrom, R. T., & Hays, G. C. (2014). The accuracy of Fastloc GPS locations and implications for animal tracking. Methods in Ecology and Evolution, 5(11), 1162–1169. https://doi.org/10.1111/2041-210X.12286
- Erbe, C., & Farmer, D. M. (2000). Zones of impact around icebreakers affecting beluga whales in the Beaufort Sea. Journal of the Acoustical Society of America, 108(3), 1332–1340. https://doi.org/10.1121/1.1288938
- Fahlman, A., Moore, M. J., & Wells, R. S. (2021). How do marine mammals manage and usually avoid gas emboli formation and gas embolic pathology? Critical clues from studies of wild dolphins. *Frontiers in Marine Science*, 8, Article 598633. https://doi.org/10.3389/fmars.2021.598633
- Finley, K. J., Miller, G. W., Davis, R. A., & Greene, C. R., Jr. (1990). Reactions of belugas (Delphinapterus leucas) and narwhals (Monodon monoceros) to ice-breaking ships in the Canadian High Arctic. Canadian Bulletin of Fisheries and Aquatic Sciences, 224, 97–117.
- Fisheries Joint Management Committee. (2013). Beaufort Sea beluga management plan [4th amended printing]. Fisheries Joint Management Committee, Inuvik, NWT, Canada.
- Ford, J. K. B., & Reeves, R. R. (2008). Fight or flight: antipredator strategies of baleen whales. Mammal Review, 38(1), 50–86. https://doi.org/10.1111/j.1365-2907.2008.00118.x
- Fraker, M. A. (1977). The 1976 white whale monitoring program, Mackenzie Estuary, NWT. Prepared for Imperial Oil Limited, Calgary, Canada, by F. F. Slaney & Company Limited, Vancouver, Canada.
- Fraker, M. A. (1978). The 1978 whale monitoring program Mackenzie Estuary, NWT. F. F. Slaney and Company Limited Report for Imperial Oil Limited, Calgary, Canada.
- Francois, R., & Garrison, G. (1982). Sound absorption based on ocean measurements: Part I: Pure water and magnesium sulfate contributions. Journal of the Acoustical Society of America, 72(3), 896–907. https://doi.org/10.1121/1.388170
- Freitas, C. (2012). argosfilter: Argos locations filter (R package version 0.63) [Computer software]. https://CRAN.Rproject.org/package=argosfilter
- Freitas, C., Lydersen, C., Fedak, M. A., & Kovacs, K. M. (2008). A simple new algorithm to filter marine mammal Argos locations. Marine Mammal Science, 24(2), 315–325. https://doi.org/10.1111/j.1748-7692.2007.00180.x
- Frost, K. J., & Suydam, R. S. (2010). Subsistence harvest of beluga or white whales (Delphinapterus leucas) in northern and western Alaska, 1987–2006. Journal of Cetacean Research and Management, 11(3), 293–299.
- Gabrielsen, G., & Smith, E. (1995). Physiological responses of wildlife to disturbance. In R. L. Knight & K. J. Gutzwiller (Eds.), Wildlife and recreationists. Coexistence through management and research (pp. 95–107). Island Press.

- GEBCO Bathymetric Compilation Group. (2020). The GEBCO_2020 Grid a continuous terrain model of the global oceans and land. British Oceanographic Data Centre, National Oceanography Centre, NERC, UK. https://doi.org/10.5285/ a29c5465-b138-234d-e053-6c86abc040b9
- Gomez, C., Lawson, J., Wright, A. J., Buren, A., Tollit, D., & Lesage, V. (2016). A systematic review on the behavioural responses of wild marine mammals to noise: the disparity between science and policy. *Canadian Journal of Zoology*, 94(12), 801–819. https://doi.org/10.1139/cjz-2016-0098
- Götz, T., Hastie, G., Hatch, L., Raustein, O., Southall, B., Tasker, M., & Thomsen, F. (2009). Overview of the impacts of anthropogenic underwater sound in the marine environment. OSPAR Biodiversity Series, 441, 1–134.
- Halliday, W. D., Barclay, D., Barkley, A. N., Cook, E., Dawson, J., Hilliard, R. C., Hussey, N. E., Jones, J. M., Juanes, F., Marcoux, M., Niemi, A., Nudds, S., Pine, M. K., Richards, C., Scharffenberg, K., Westdal, K., & Insley, S. J. (2021). Underwater sound levels in the Canadian Arctic, 2014–2019. *Marine Pollution Bulletin*, 168, Article 112437. https://doi.org/ 10.1016/j.marpolbul.2021.112437, 112437
- Halliday, W. D., Insley, S. J., Hilliard, R. C., de Jong, T., & Pine, M. K. (2017). Potential impacts of shipping noise on marine mammals in the western Canadian Arctic. *Marine Pollution Bulletin*, 123(1–2), 73–82. https://doi.org/10.1016/ j.marpolbul.2017.09.027
- Halliday, W. D., Pine, M. K., Citta, J. J., Harwood, L., Hauser, D. D. W., Hilliard, R. C., Lea, E. V., Loseto, L. L., Quakenbush, L., & Insley, S. J. (2021). Potential exposure of beluga and bowhead whales to underwater noise from ship traffic in the Beaufort and Chukchi Seas. Ocean & Coastal Management, 204, Article 105473. https://doi.org/10.1016/ j.ocecoaman.2020.105473
- Halliday, W. D., Pine, M. K., & Insley, S. J. (2020a). Underwater noise and Arctic marine mammals: review and policy recommendations. *Environmental Reviews*, 28(4), 438–448. https://doi.org/10.1139/er-2019-0033
- Halliday, W. D., Pine, M. K., Mouy, X., Kortsalo, P., Hilliard, R. C., & Insley, S. J. (2020b). The coastal Arctic marine soundscape near Ulukhaktok, Northwest Territories, Canada. *Polar Biology*, 43(6), 623–636. https://doi.org/10.1007/s00300-020-02665-8
- Halliday, W. D., Scharffenberg, K., MacPhee, S., Hilliard, R. C., Mouy, X., Whalen, D., Loseto, L. L., & Insley, S. J. (2019). Beluga vocalizations decrease in response to vessel traffic in the Mackenzie River estuary. Arctic, 72(4), 337–346. https://doi.org/10.14430/arctic69294
- Halliday, W. D., Têtu, P.-L., Dawson, J., Insley, S. J., & Hilliard, R. C. (2018). Tourist vessel traffic in important whale areas in the western Canadian Arctic: Risks and possible management solutions. *Marine Policy*, 97, 72–81. https://doi.org/ 10.1016/j.marpol.2018.08.035
- Harwood, L. A., Iacozza, J., Auld, J. C., Norton, P., & Loseto, L. (2014b). Belugas in the Mackenzie River estuary, NT, Canada: habitat use and hot spots in the Tarium Niryutait Marine Protected Area. Ocean & Coastal Management, 100, 128–138. https://doi.org/10.1016/j.ocecoaman.2014.08.004
- Harwood, L. A., Kingsley, M. C., & Smith, T. G. (2014a). An emerging pattern of declining growth rates in belugas of the Beaufort Sea: 1989–2008. Arctic, 67, 483–492. https://doi.org/10.14430/arctic4423
- Harwood, L., & Smith, T. (2002). Beaufort Sea whales: an overview and outlook. Arctic, 55(Suppl 1), 77-93.
- Hauser, D. D. W., Laidre, K. L., & Stern, H. L. (2018). Vulnerability of Arctic marine mammals to vessel traffic in the increasingly ice-free Northwest Passage and Northern Sea Route. Proceedings of the National Academy of Sciences of the United States of America, 115(29), 7617–7622. https://doi.org/10.1073/pnas.1803543115
- Higdon, J., Byers, T., Brown, L., & Ferguson, S. (2013). Observations of killer whales (Orcinus orca) in the Canadian Beaufort Sea. Polar Record, 49(3), 307–314. https://doi.org/10.1017/S0032247412000356
- Hill, H., Dietrich, S., Yeater, D., McKinnon, M., Miller, M., Aibel, S., & Dove, A. (2015). Developing a catalog of socio-sexual behaviors of beluga whales (Delphinapterus leucas). Animal Behavior and Cognition, 2(2), 105–123. https://doi.org/ 10.12966/abc.05.01.2015
- Hobbs, R. C., Shelden, K. E. W., Vos, D. J., Goetz, K. T., & Rugh, D. J. (2006). Status review and extinction assessment of Cook Inlet belugas (Delphinapterus leucas) (AFSC Processed Report 2006–16). Alaska Fisheries Science Center, NOAA, National Marine Fisheries Services.
- Hooker, S. K., Baird, R. W., & Fahlman, A. (2009). Could beaked whales get the bends? Effect of diving behaviour and physiology on modelled gas exchange for three species: *Ziphius cavirostris*, *Mesoplodon densirostris* and *Hyperoodon ampullatus*. *Respiratory Physiology & Neurobiology*, 167(3), 235–246. https://doi.org/10.1016/j.resp.2009.04.023
- Hornby, C. A., Hoover, C., Iacozza, J., Barber, D. G., & Loseto, L. L. (2016). Spring conditions and habitat use of beluga whales (*Delphinapterus leucas*) during arrival to the Mackenzie River Estuary. *Polar Biology*, 39(12), 2319–2334. https://doi.org/ 10.1007/s00300-016-1899-9
- Huntington, H. P. (2000). Traditional knowledge of the ecology of belugas, *Delphinapterus leucas*, in Cook Inlet, Alaska. *Marine Fisheries Review*, 62(3), 134–140.
- Hussey, N. E., Kessel, S. T., Aarestrup, K., Cooke, S. J., Cowley, P. D., Fisk, A. T., Harcourt, R. G., Holland, K. N., Iverson, S. J., Kocik, J. F., Mills Flemming, J. E., & Whoriskey, F. G. (2015). Aquatic animal telemetry: a panoramic window into the underwater world. *Science*, 348(6240), Article 1255642. https://doi.org/10.1126/science.1255642

- Insley, S. J., Halliday, W. D., & de Jong, T. (2017). Seasonal patterns in ocean ambient noise near Sachs Harbour, Northwest Territories. Arctic, 70, 239–248. https://doi.org/10.14430/arctic4662
- Intergovernmental Panel on Climate Change. (2022). Climate Change 2022: Impacts, adaptation and vulnerability. *IPCC WGII Sixth* Assessment Report. https://www.ipcc.ch/report/ar6/wg2/downloads/report/IPCC_AR6_WGII_FinalDraft_FullReport.pdf

International Maritime Organization. (2014). International convention for the safety of life at sea (SOLAS).

- Irvine, L., Palacios, D. M., Urbán, J., & Mate, B. (2017). Sperm whale dive behavior characteristics derived from intermediate duration archival tag data. *Ecology and Evolution*, 7(19), 7822–7837. https://doi.org/10.1002/ece3.3322
- Jakobsson, M., Mayer, L., Coakley, B., Dowdeswell, J. A., Forbes, S., Fridman, B., Hodnesdal, H., Noormets, R., Pedersen, R., Rebesco, M., Schenke, H. W., Zarayskaya, Y., Accettella, D., Armstrong, A., Anderson, R. M., Bienhoff, P., Camerlenghi, A., Church, I., Edwards, M., ... Weatherall, P. (2012). The International Bathymetric Chart of the Arctic Ocean (IBCAO) version 3.0. *Geophysical Research Letters*, 39(12). https://doi.org/10.1029/2012GL052219
- Jefferson, T. A., Leatherwood, S., & Webber, M. A. (1993). Marine mammals of the world. Food and Agriculture Organization of the United Nations.
- Johnson, C., McManus, M., & Skaar, D. (1989). Masked tonal hearing thresholds in the beluga whale. Journal of the Acoustical Society of America, 85(6), 2651–2654. https://doi.org/10.1121/1.397759
- Johnson, D. S., & London, J. M. (2018). R package 'crawl' (R package version 2.2/1) [Computer software]. https://cran.rproject.org/web/packages/crawl/
- Johnson, D. S., London, J. M., Lea, M. A., & Durban, J. W. (2008). Continuous time correlated random walk model for animal telemetry data. *Ecology*, 89(5), 1208–1215. https://doi.org/10.1890/07-1032.1
- Kinda, G. B., Simard, Y., Gervaise, C., Mars, J. I., & Fortier, L. (2013). Under-ice ambient noise in Eastern Beaufort Sea, Canadian Arctic, and its relation to environmental forcing. *Journal of the Acoustical Society of America*, 134(1), 77–87. https:// doi.org/10.1121/1.4808330
- Koski, W. R., Miller, G. W., Patenaude, N. J., Richardson, W. J., & Smultea, M. A. (1995). White whale results. In Acoustic effects of oil production activities on bowhead and white whales visible during spring migration near Pt. Barrow, Alaska – 1991 and 1994 phases (LGL Report TA954). U.S. Minerals Management Service (pp. 338–386). https://www.boem.gov/ sites/default/files/boem-newsroom/Library/Publications/1995/95_0051.pdf
- Krasnova, V. V., Bel'kovich, V. M., & Chernetskii, A. D. (2009). Formation of behavior in the White Sea beluga calf, Delphinapterus leucas, during early postnatal ontogenesis. Russian Journal of Marine Biology, 35(1), 53–59. https:// doi.org/10.1134/s1063074009010088
- Lemieux Lefebvre, S., Lesage, V., Michaud, R., & Humphries, M. (2018). Classifying and combining herd surface activities and individual dive profiles to identify summer behaviours of beluga (*Delphinapterus leucas*) from the St. Lawrence Estuary, Canada. *Canadian Journal of Zoology*, 96(5), 393–410. https://doi.org/10.1139/cjz-2017-0015
- Lenth, R. V. (2016). Least-squares means: The R package Ismeans. Journal of Statistical Software, 69(1), 1–33. https:// doi.org/10.18637/jss.v069.i01
- LGL & Greeneridge. (1986). Reactions of beluga whales and narwhals to ship traffic and ice-breaking along ice edges in the Eastern Canadian High Arctic: 1982–1984. Environmental Studies 37. Indian and Northern Affairs Canada.
- Loseto, L., Richard, P., Stern, G., Orr, J., & Ferguson, S. (2006). Segregation of Beaufort Sea beluga whales during the openwater season. *Canadian Journal of Zoology*, 84(12), 1743–1751. https://doi.org/10.1139/Z06-160
- MacGillivray, A., & de Jong, C. A. (2021). Reference spectrum model for estimating source levels of marine shipping based on automated identification system data. *Journal of Marine Science and Engineering*, 9, Article 369. https://doi.org/ 10.3390/jmse9040369
- Martin, A., Kingsley, M., & Ramsay, M. (1994). Diving behaviour of narwhals (Monodon monoceros) on their summer grounds. Canadian Journal of Zoology, 72(1), 118–125. https://doi.org/10.1139/z94-015
- Martin, A. R., Smith, T. G., & Cox, O. P. (1998). Dive form and function in belugas *Delphinapterus leucas* of the eastern Canadian High Arctic. *Polar Biology*, 20(3), 218–228. https://doi.org/10.1007/s003000050299
- McCarty, S. (1981). Survey of effects of outer continental shelf platforms on cetacean behavior. App. C. Vol. II. Effects of noise of offshore oil and gas operations on marine mammals. An introductory assessment. NOSC Technical Report 844.
- Meredith, M., Sommerkorn, M., Cassotta, S., Derksen, C., Ekaykin, A., Hollowed, A., Kofinas, G., Mackintosh, A., Melbourne-Thomas, J., Muelbert, M. M. C., Ottersen, G., Pritchard, H., & Schuur, E. A. G. (2019). Polar regions. In H.-O. Pörtner, D. C. Roberts, V. Masson-Delmotte, P. Zhai, M. Tignor, E. Poloczanska, K. Mintenbeck, A. Alegría, M. Nicolai, A. Okem, J. Petzold, B. Rama, & N. M. Weyer (Eds.), *IPCC Special Report on the ocean and cryosphere in a changing climate* (pp. 203–320). Cambridge University Press. https://doi.org/10.1017/9781009157964.005
- Miller, P. J., Kvadsheim, P. H., Lam, F. P., Tyack, P. L., Cure, C., DeRuiter, S. L., Kleivane, L., Sivle, L. D., van IJsselmuide, S. P., Visser, F., Wensveen, P. J., von Benda-Beckmann, A. M., Martin Lopez, L. M., Narazaki, T., & Hooker, S. K. (2015). First indications that northern bottlenose whales are sensitive to behavioural disturbance from anthropogenic noise. *Royal Society Open Science*, 2(6), Article 140484. https://doi.org/10.1098/rsos.140484

- Miller, G. W., Moulton, V. D., Davis, R. A., Holst, M., Millman, P., MacGillivray, A., & Hannay, D. (2005). Monitoring seismic effects on marine mammals—southeastern Beaufort Sea, 2001–2002. In Offshore oil and gas environmental effects monitoring: Approaches and technologies (pp. 511–542). Battelle Press.
- Mooney, T. A., Castellote, M., Jones, I., Rouse, N., Rowles, T., Mahoney, B., & Goertz, C. E. C. (2020). Audiogram of a Cook Inlet beluga whale (Delphinapterus leucas). Journal of the Acoustical Society of America, 148(5), Article 3141. https:// doi.org/10.1121/10.0002351, 3148
- Mooney, T. A., Castellote, M., Quakenbush, L., Hobbs, R., Gaglione, E., & Goertz, C. (2018). Variation in hearing within a wild population of beluga whales (*Delphinapterus leucas*). Journal of Experimental Biology, 221(Pt 9), Article jeb171959. https://doi.org/10.1242/jeb.171959
- Mudryk, L. R., Dawson, J., Howell, S. E., Derksen, C., Zagon, T. A., & Brady, M. (2021). Impact of 1, 2 and 4°C of global warming on ship navigation in the Canadian Arctic. *Nature Climate Change*, 11, 673–679. https://doi.org/10.1038/ s41558-021-01087-6
- National Marine Fisheries Service. (2018). Revisions to: Technical guidance for assessing the effects of anthropogenic sound on marine mammal hearing (version 2.0): underwater thresholds for onset of permanent and temporary threshold shifts (NOAA Technical Memorandum NMFS-OPR-59). U.S. Department of Commerce.
- New, L., Lusseau, D., & Harcourt, R. (2020). Dolphins and boats: when is a disturbance, disturbing? Frontiers in Marine Science, 7, Article 353. https://doi.org/10.3389/fmars.2020.00353
- Nunes, A. (2019). Divebomb dive classification algorithm (Python package version 1.1.2) [Computer software]. https://pypi.org/project/divebomb/
- Orr, J. R., Joe, R., & Evic, D. (2001). Capturing and handling of white whales (*Delphinapterus leucas*) in the Canadian Arctic for instrumentation and release. Arctic, 54(3), 299–304. https://doi.org/10.14430/arctic789
- Palomino González, A., Kovacs, K. M., Lydersen, C., Ims, R. A., & Lowther, A. D. (2021). Drones and marine mammals in Svalbard, Norway. Marine Mammal Science, 37(4), 1212–1229. https://doi.org/10.1111/mms.12802
- Patenaude, N. J., Richardson, W. J., Smultea, M. A., Koski, W. R., Miller, G. W., Würsig, B., & Greene, C. R. (2002). Aircraft sound and disturbance to bowhead and beluga whales during spring migration in the Alaskan Beaufort Sea. *Marine Mammal Science*, 18(2), 309–335. https://doi.org/10.1111/j.1748-7692.2002.tb01040.x
- Payne, R., & Webb, D. (1971). Orientation by means of long range acoustic signaling in baleen whales. Annals of the New York Academy of Sciences, 188(1), 110–141. https://doi.org/10.1111/j.1749-6632.1971.tb13093.x
- Pine, M. K., Hannay, D. E., Insley, S. J., Halliday, W. D., & Juanes, F. (2018). Assessing vessel slowdown for reducing auditory masking for marine mammals and fish of the western Canadian Arctic. *Marine Pollution Bulletin*, 135, 290–302. https:// doi.org/10.1016/j.marpolbul.2018.07.031
- Pirotta, E., Merchant, N. D., Thompson, P. M., Barton, T. R., & Lusseau, D. (2015). Quantifying the effect of boat disturbance on bottlenose dolphin foraging activity. *Biological Conservation*, 181, 82–89. https://doi.org/10.1016/j.biocon.2014.11.003
- Popov, V. V., Supin, A. Y., Rozhnov, V. V., Nechaev, D. I., Sysuyeva, E. V., Klishin, V. O., Pletenko, M. G., & Tarakanov, M. B. (2013). Hearing threshold shifts and recovery after noise exposure in beluga whales, *Delphinapterus leucas. Journal of Experimental Biology*, 216(Pt 9), 1587–1596. https://doi.org/10.1242/jeb.078345
- Protection of the Arctic Marine Environment. (2019). Underwater noise in the Arctic: A state of knowledge report. PAME Secretariat, Akureyri, Iceland.
- Quick, N. J., Isojunno, S., Sadykova, D., Bowers, M., Nowacek, D. P., & Read, A. J. (2017). Hidden Markov models reveal complexity in the diving behaviour of short-finned pilot whales. *Scientific Reports*, 7(1), Article 45765. https://doi.org/ 10.1038/srep45765.
- R Core Team. (2021). R: a language and environment for statistical computing [Computer software]. R Foundation for Statistical Computing.
- Reeves, R., Rosa, C., George, J. C., Sheffield, G., & Moore, M. (2012). Implications of Arctic industrial growth and strategies to mitigate future vessel and fishing gear impacts on bowhead whales. *Marine Policy*, 36(2), 454–462. https://doi.org/ 10.1016/j.marpol.2011.08.005
- Richard, P. R., Martin, A. R., & Orr, J. R. (2001). Summer and autumn movements of belugas of the eastern Beaufort Sea stock. Arctic, 54(3), 223–236. https://doi.org/10.14430/arctic783
- Richardson, W. J., Greene, C. R., Jr., Koski, W. R., Malme, C. I., Miller, G. W., Smultea, M. A., & Würsig, B. (1990). Acoustic effects of oil-production activities on bowhead and white whales visible during spring migration near Pt. Barrow, Alaska-1989 phase: Sound propagation and whale responses to playbacks of continuous drilling noise from an ice platform, as studied in pack ice conditions (LGL Report TA848-4). U.S. Minerals Management Service. https://www.boem.gov/sites/default/ files/boem-newsroom/Library/Publications/1990/90_0017.pdf
- Richardson, W. J., Greene, C. R., Jr., Malme, C. I., & Thomson, D. H. (1995). Marine mammals and noise. Academic Press.
- Ridgway, S. H., Carder, D. A., Kamolnick, T., Smith, R. R., Schlundt, C. E., & Elsberry, W. R. (2001). Hearing and whistling in the deep sea: depth influences whistle spectra but does not attenuate hearing by white whales (*Delphinapterus leucas*) (Odontoceti, Cetacea). *Journal of Experimental Biology*, 204(22), 3829–3841. https://doi.org/10.1242/jeb.204.22.3829

- Roelofs, K. (2017). Freeze for action: neurobiological mechanisms in animal and human freezing. Philosophical Transactions of the Royal Society B: Biological Sciences, 372(1718), Article 20160206. https://doi.org/10.1098/rstb.2016.0206
- Roth, E. H., Hildebrand, J. A., Wiggins, S. M., & Ross, D. (2012). Underwater ambient noise on the Chukchi Sea continental slope from 2006–2009. Journal of the Acoustical Society of America, 131(1), 104–110. https://doi.org/10.1121/1.3664096
- Schack, H., & Haapaniemi, J. (2017). Potential impact of noise from shipping on key species of marine mammals in waters off Western Greenland-Case Baffinland (Report Number 30-06-2017). WWF Denmark, Copenhagen, Denmark.
- Simon, M., Johnson, M., Tyack, P., & Madsen, P. T. (2009). Behaviour and kinematics of continuous ram filtration in bowhead whales (Balaena mysticetus). Proceedings of the Royal Society B: Biological Sciences, 276(1674), 3819–3828. https:// doi.org/10.1098/rspb.2009.1135
- Southall, B. L., Bowles, A. E., Ellison, W. T., Finneran, J., Gentry, R. L., Greene Jr, C. R., Kastak, D., Ketten, D. R., Miller, J. H., Nachtigall, P. E., Richardson, W. J., & Tyack, P. L. (2007). Marine mammal noise exposure criteria: initial scientific recommendations. Aquatic Mammals, 33(4), 427–436.
- Southall, B. L., Finneran, J. J., Reichmuth, C., Nachtigall, P. E., Ketten, D. R., Bowles, A. E., Ellison, W. T., Nowacek, D. P., & Tyack, P. L. (2019). Marine mammal noise exposure criteria: updated scientific recommendations for residual hearing effects. Aquatic Mammals, 45(2), 125–232. https://doi.org/10.1578/am.45.2.2019.125
- Spiess, A.-N. (2014). *qpcR: Modelling and analysis of real-time PCR data* (R package version 4.1.2) [Computer software]. https://rdrr.io/cran/qpcR/
- Stafford, K. M. (2019). Increasing detections of killer whales (Orcinus orca), in the Pacific Arctic. Marine Mammal Science, 35(2), 696–706. https://doi.org/10.1111/mms.12551
- Stanek, R. (1996). Belukha hunters of Cook Inlet, Alaska. In N. Davis (Ed.), Proceedings of the Symposium of the Cook Inlet Historical Society.
- Steckenreuter, A., Harcourt, R., & Möller, L. (2011). Distance does matter: close approaches by boats impede feeding and resting behaviour of Indo-Pacific bottlenose dolphins. Wildlife Research, 38(6), 455–463. https://doi.org/10.1071/ WR11048
- Steen, J., Gabrielsen, G., & Kanwisher, J. (1988). Physiological aspects of freezing behaviour in willow ptarmigan hens. Acta Physiologica Scandinavica, 134(2), 299–304. https://doi.org/10.1111/j.1748-1716.1988.tb08493.x
- Stewart, B. S., Awbrey, F. T., & Evans, W. E. (1983). Belukha whale (Delphinapterus leucas) responses to industrial noise in Nushagak Bay, Alaska: 1983. NOAA, OCSEAP Final Report 43(1986) (pp. 587–616). U.S. Department of Commerce.
- Stewart, B., Evans, W., & Awbrey, F. (1982). Effects of man-made waterborne noise on behavior of belukha whales (Delphinapterus leucas) in Bristol Bay, Alaska (Unpublished report for National Oceanic and Atmospheric Administration, Juneau, Alaska). Hubbs/Sea World Research Institute Technical Report 82–145.
- Storrie, L., Hussey, N. E., MacPhee, S. A., O'Corry-Crowe, G., Iacozza, J., Barber, D. G., Nunes, A., & Loseto, L. L. (2022). Year-round dive characteristics of male beluga whales from the eastern Beaufort Sea population indicate seasonal shifts in foraging strategies. *Frontiers in Marine Science*, 8, Article 715412. https://doi.org/10.3389/fmars.2021.715412
- Tyack, P. (1986). Population biology, social behavior and communication in whales and dolphins. *Trends in Ecology & Evolution*, 1(6), 144–150. https://doi.org/10.1016/0169-5347(86)90042-X
- Tyack, P. L. (2009). Human-generated sound and marine mammals. Physics Today, 62(11), 39–44. https://doi.org/10.1063/ 1.3265235
- Tyack, P. L., Zimmer, W. M. X., Moretti, D., Southall, B. L., Claridge, D. E., Durban, J. W., Clark, C. W., D'Amico, A., DiMarzio, N., Jarvis, S., McCarthy, E., Morrissey, R., Ward, J., & Boyd, I. L. (2011). Beaked whales respond to simulated and actual navy sonar. *PloS ONE*, 6(3), Article e17009. https://doi.org/10.1371/journal.pone.0017009, 6
- Urick, R. J. (Ed.) (1983). Principles of underwater sound (3rd ed.). Peninsula Publishing.
- Vacquié-Garcia, J., Lydersen, C., & Kovacs, K. M. (2019). Diving behaviour of adult male white whales (Delphinapterus leucas) in Svalbard, Norway. Polar Research, 38, Article 3605. https://doi.org/10.33265/polar.v38.3605
- Van Rossum, G., & Drake, F. (2009). Python 3 reference manual. CreateSpace.
- Veirs, S., Veirs, V., & Wood, J. D. (2016). Ship noise extends to frequencies used for echolocation by endangered killer whales. PeerJ, 4, Article e1657. https://doi.org10.7717/peerj.1657
- Vergara, V., Wood, J., Lesage, V., Ames, A., Mikus, M.-A., & Michaud, R. (2021). Can you hear me? Impacts of underwater noise on communication space of adult, sub-adult and calf contact calls of endangered St. Lawrence belugas (Delphinapterus leucas). Polar Research, 40, 1–19. https://doi.org/10.33265/polar.v40.5521
- Vincent, C., Mcconnell, B. J., Ridoux, V., & Fedak, M. A. (2002). Assessment of Argos location accuracy from satellite tags deployed on captive gray seals. *Marine Mammal Science*, 18(1), 156–166. https://doi.org/10.1111/j.1748-7692.2002.tb01025.x
- White, M. J., Jr., Norris, J., Ljungblad, D., Baron, K., & di Sciara, G. (1978). Auditory threshold of two beluga whales (Delphinapterus leucas). Hubbs/Sea World Research Institute Technical Report No. 78–109.
- Whitehead, H. (2016). Consensus movements by groups of sperm whales. *Marine Mammal Science*, 32(4), 1402–1415. https://doi.org/10.1111/mms.12338

- Williams, T. M., Blackwell, S. B., Tervo, O., Garde, E., Sinding, M. H. S., Richter, B., & Heide Jørgensen, M. P. (2022). Physiological responses of narwhals to anthropogenic noise: A case study with seismic airguns and vessel traffic in the Arctic. *Functional Ecology*, 36(9), 2251–2266. https://doi.org/10.1111/1365-2435.14119
- Williams, R., Cholewiak, D., Clark, C. W., Erbe, C., George, C., Lacy, R., Leaper, R., Moore, S., New, L., Parsons, C., Rosenbaum, H., Rowles, T., Simmonds, M., Stimmelmayr, R., Suydam, R. S., & Wright, A. (2020). Chronic ocean noise and cetacean population models. *Journal of Cetacean Research and Management*, 21(1), 85–94. https://doi.org/10.47536/ jcrm.v21i1.202
- Wisniewska, D. M., Johnson, M., Teilmann, J., Siebert, U., Galatius, A., Dietz, R., & Madsen, P. T. (2018). High rates of vessel noise disrupt foraging in wild harbour porpoises (*Phocoena phocoena*). Proceedings of the Royal Society B: Biological Sciences, 285(1872), Article 20172314. https://doi.org/10.1098/rspb.2017.2314
- Wood, S. & Scheipl, F. (2020). gamm4: Generalized Additive Mixed Models using 'mgcv' and 'lme4' (R package version 0.2-6) [Computer software]. https://CRAN.R-project.org/package=gamm4
- Yang, T. C., & Votaw, C. W. (1981). Under ice reflectivities at frequencies below 1 kHz. Journal of the Acoustical Society of America, 70(3), 841–851. https://doi.org/10.1121/1.386924
- Zeng, Q., Lu, T., Lin, K.-C., Yuen, K. F., & Li, K. X. (2020). The competitiveness of Arctic shipping over Suez Canal and China-Europe railway. Transport Policy, 86, 34–43. https://doi.org/10.1016/j.tranpol.2019.11.005

SUPPORTING INFORMATION

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